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# 1 Distinct anaerobic bacterial consumers of cellobiose-derived carbon in boreal fens

# 2 differing in CH<sub>4</sub> vs. CO<sub>2</sub> production ratio

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# 11

12 Running title: Anaerobic cellobiose consumers in boreal fens

# 13

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18

# 19 Abstract

20 Northern peatlands in general have high methane (CH<sub>4</sub>) emissions, but individual peatlands show 21 considerable variation as CH<sub>4</sub> sources. Particularly in nutrient poor peatlands, CH<sub>4</sub> production can be 22 low and exceeded by carbon dioxide (CO<sub>2</sub>) production from unresolved anaerobic processes. To clarify 23 the role anaerobic bacterial degraders play in this variation, we compared consumers of cellobiose-24 derived carbon in two fens differing in nutrient status and the ratio of CH<sub>4</sub> and CO<sub>2</sub> produced. After  $[^{13}C]$  cellobiose amendment, the mesotrophic fen produced equal amounts of CH<sub>4</sub> and CO<sub>2</sub>. The 25 26 oligotrophic fen had lower CH<sub>4</sub> production but produced 3-59 times more CO<sub>2</sub> than CH<sub>4</sub>. RNA stable 27 isotope probing revealed that in the mesotrophic fen with higher CH<sub>4</sub> production cellobiose-derived carbon was mainly assimilated by various recognized fermenters of Firmicutes and by Proteobacteria. 28 29 The oligotrophic peat with excess CO<sub>2</sub> production revealed a wider variety of cellobiose-C consumers 30 including Firmicutes and Proteobacteria but additionally more unconventional degraders, such as 31 Telmatobacter-related Acidobacteria and subphylum 3 of Verrucomicrobia. Prominent and potentially 32 fermentative Planctomycetes and Chloroflexi did not appear to process cellobiose-C. Our results show 33 that anaerobic degradation resulting in different levels of CH<sub>4</sub> production can involve distinct sets of 34 bacterial degraders. By distinguishing cellobiose degraders from the total community, this study 35 contributes to defining anaerobic bacteria that process cellulose-derived carbon in peat. Several of the identified degraders, particularly fermenters and potential Fe(III) or humic substance reducers in the 36 37 oligotrophic peat, represent promising candidates for resolving the origin of excess CO<sub>2</sub> production in 38 peatlands.

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# 39 Importance

| 40 | Peatlands are major sources of the greenhouse gas methane (CH <sub>4</sub> ), yet in many peatlands CO <sub>2</sub>      |
|----|--|
| 41 | production from unresolved anaerobic processes exceeds CH4 production. Anaerobic degradation                             |
| 42 | produces the precursors of $CH_4$ production but also represents competing processes. We show that                       |
| 43 | anaerobic degradation leading to high or low CH4 production involved distinct sets of bacteria. Well-                    |
| 44 | known fermenters dominated in a peatland with high CH <sub>4</sub> production, while novel and unconventional            |
| 45 | degraders could be identified in a site where CO <sub>2</sub> production greatly exceeds CH <sub>4</sub> production. Our |
| 46 | results help identify and assign functions to uncharacterized bacteria that promote or inhibit $CH_4$                    |
| 47 | production and reveal bacteria potentially producing the excess CO <sub>2</sub> in acidic peat. This study               |
| 48 | contributes to understanding the microbiological basis for different levels of CH <sub>4</sub> emission from             |
| 49 | peatlands.   |

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# 50 Introduction

Methane (CH<sub>4</sub>) emission rates from peatlands, which together with other wetlands form the largest natural source of CH<sub>4</sub> (1), vary considerably with water table level, temperature, vegetation, and peatland type (2, 3). These factors have been shown to affect the activity and community composition of CH<sub>4</sub>-producing archaea and in some cases the pathway of CH<sub>4</sub> production (4-11). It is much less clear how the environmental variation of peatland ecosystems influences the anaerobic bacterial degraders, which produce the precursors of methanogenesis and on the other hand compete with methanogens for substrates.

58 Peat is partially decomposed plant material, the main components being cellulose, 59 hemicellulose, and recalcitrant compounds such as lignin and humic substances (12). In anoxic peat 60 below the water level, decomposition of plant-derived polysaccharides requires a cooperation of 61 functional microbial groups: primary fermenters hydrolyzing polymers and fermenting the monomers such as sugars, secondary fermenters turning the resulting organic acids into the methanogenic 62 63 substrates acetate and  $H_2$ +CO<sub>2</sub>, and finally methanogens producing CH<sub>4</sub> (13). Polysaccharide 64 hydrolysis and thus the activity of non-methanogenic anaerobic bacteria has been proposed to be the 65 rate-limiting step for wetland CH<sub>4</sub> production (5, 14-16).

The complete anaerobic degradation of carbohydrates under methanogenic conditions, when both acetate and  $H_2+CO_2$  are precursors of methanogenesis, should theoretically produce equal amounts of  $CH_4$  and  $CO_2$  (17). In contrast, a large excess of anaerobic  $CO_2$  production is often observed in peat, particularly in low pH and nutrient-poor peatlands (18-22). These sites in particular but northern peatlands in general have low levels of Fe(III), nitrate, and sulfate, which is considered to limit  $CO_2$  production resulting from the degradation of organic matter with the reduction of these inorganic electron acceptors (18, 19). Small sulfate pools in peat may have unexpectedly high turnover

73 rates (23, 24), but currently the main proposed sources for the excess  $CO_2$  in nutrient-poor peat are 74 fermentation or anaerobic respiration with phenolic and quinone-containing compounds of organic 75 residues, commonly referred to as humic substances, as electron acceptors (13, 19, 21, 25). 76 Identification of the anaerobic bacterial degraders active in peatlands with contrasting CH<sub>4</sub> vs. CO<sub>2</sub> 77 production could provide insights into the processes producing the excess CO<sub>2</sub>. The current information on peat microbes involved in the intermediary anaerobic 78 79 processes, leading to methanogenesis or excess CO<sub>2</sub> production, is still restricted to relatively few 80 studies and sites. Work on a temperate spruce fen has identified cellulose, glucose, and xylose 81 fermenters, including Bacteroidetes, Acidobacteria, Firmicutes, Spirochaeta, Actinobacteria, and 82 Proteobacteria (26-30). Metagenomic studies in a Sphagnum bog and in Arctic peat have indicated Applied and Environmental 83 Firmicutes, Actinobacteria, Bacteroidetes, and Deltaproteobacteria as potential fermenters (14, 31, Microbioloav 32). A comparison of eutrophic to oligotrophic temperate peatlands found indications of functional 84

85 redundancy, with similar anaerobic processes being carried out by different communities (33). This

86 raises the question on whether and how the community composition of the anaerobic bacterial

87 degraders relates to the eventual CH<sub>4</sub> production of a peatland.

88 We set out to compare the active anaerobic bacterial degraders in two boreal peatlands 89 differing as CH<sub>4</sub> sources. These adjacent Sphagnum fens are rather similar in vegetation composition 90 and pH but differ in nutrient status and, more importantly, in the pathway of organic matter degradation 91 and the ratio of  $CH_4$  and  $CO_2$  produced (21). The mesotrophic fen has higher  $CH_4$  production, higher 92 contribution of the acetoclastic pathway of CH<sub>4</sub> production and a distinct methanogen community from 93 the oligotrophic fen (8, 21, 34). We targeted consumers of cellobiose-derived carbon, representing 94 bacteria processing labile carbon from cellulose degradation, by incubating peat with  $[^{13}C]$  cellobiose 95 and following the label in the produced  $CH_4$  and  $CO_2$  and in RNA by stable isotope probing (SIP). Our

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aim was to 1) identify cellobiose consumers and fermenters in boreal *Sphagnum* fens and 2) compare
the active anaerobic degraders in two fens differing as CH<sub>4</sub> sources. By selecting outwardly similar but
functionally contrasting sites, we aimed to differentiate between degraders active under high CH<sub>4</sub>
production and under lower CH<sub>4</sub> production when CO<sub>2</sub> is the main gaseous end product.

100

# 101 Materials and methods

# 102 Study site and sampling

103 We sampled an oligotrophic fen and a mesotrophic fen within the Lakkasuo mire complex in southern 104 Finland (61°47'N, 24°18'E) (35). The sampling sites are located within 500 meters from each other. 105 Both fens are minerotrophic (i.e., in contact with groundwater) with the mesotrophic fen being more 106 nutrient-rich than the oligotrophic fen (35). Vegetation at the oligotrophic site consists mainly of sedge 107 Carex lasiocarpa and mosses Sphagnum papillosum and S. fallax. The main species at the mesotrophic 108 site are sedge C. lasiocarpa and mosses S. flexuosum and S. fallax. Peat at the sampled depths consists 109 of partially degraded plant remains with no observable mineral soil. Water table depth at the time of 110 sampling in November 2011 was at 15±2 cm below peat surface at the oligotrophic site and at 4±2 cm 111 below surface at the mesotrophic site. Peat pH was  $4.8\pm0.1$  at the oligotrophic site and  $5.1\pm0.1$  at the 112 mesotrophic site.

113 Three replicate peat cores were collected from each site by cutting out a  $10 \text{ cm} \times 10 \text{ cm}$ 114 core with a knife to the depth of 30 cm from peat surface. Peat cores were stored overnight at +4 °C 115 and divided into two sections, 10-20 cm and 20-30 cm from peat surface. A portion of the sections was 116 frozen for RNA analysis of the initial peat. For initial measurements of CH<sub>4</sub> and CO<sub>2</sub> production 117 potential and for further incubation for the labeling experiment, 15 ml of peat was placed in 120-ml

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118flasks with 30 ml of  $H_2O$  previously flushed with  $N_2$  to remove oxygen. Each sample was bottled in119duplicate. Flushing with  $N_2$  was repeated, and the flasks were closed with rubber septa and incubated in120dark at 15 °C. Methane and  $CO_2$  production were followed by gas chromatography (36) with four121samplings of the headspace gas during 191 hours. Production rates include gaseous and dissolved gases122and are given per grams dry weight (gdw). Ratio of peat dry weight to peat wet weight was123 $0.099 \pm 0.007.$ 

124

# 125 Incubation and sampling for stable isotope analysis

The incubation for the labeling experiment started with addition of uniformly labeled [<sup>13</sup>C]cellobiose 126 (Omicron Biomedicals) to one of each pair of flasks in the initial measurements for the 20-30 cm layer, 127 128 when peat had been in the flasks for 10 days. Before the addition, the flasks were flushed with  $N_2$  to 129 remove accumulated gases. The flasks received 0.5 ml of 18 mM cellobiose to a final concentration of 0.2 mM injected through the septum. The other flask of each pair received the same amount of 130 <sup>12</sup>C]cellobiose as a control for <sup>13</sup>C enrichment. The flasks were shaken by hand and incubated in dark 131 at 15 °C. The cellobiose addition was repeated 12 times during 25 days (every second or third day). 132 Altogether, each flask received 118 µmol of cellobiose and 1.42 mmol of <sup>13</sup>C, yielding 94 µmol <sup>13</sup>C (or 133  $^{12}$ C for controls) per g (fresh weight) of peat. During the pulse labeling, CH<sub>4</sub> and CO<sub>2</sub> production was 134 followed with three measurements per week. The flasks were sampled for analysis of <sup>13</sup>C-CH<sub>4</sub>, <sup>13</sup>C-135 CO<sub>2</sub> and peat at 7, 14, and 28 days. Headspace gas was sampled through the septum and injected into 136 N<sub>2</sub>-flushed 12-ml vials (Labco). After gas sampling, the flasks were opened under N<sub>2</sub> flow and  $2 \times 1.5$ 137 138 ml of peat slurry was removed and frozen at -80 °C for RNA extraction for SIP.

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# 140 Isotope analysis of CO<sub>2</sub> and CH<sub>4</sub>

The <sup>13</sup>C/<sup>12</sup>C ratio of CO<sub>2</sub> and CH<sub>4</sub> was analyzed from the 12-ml vials described above. First, 1 ml was 141 injected into a second, pre-evacuated and N<sub>2</sub>-flushed 12-ml vial, which was analyzed for <sup>13</sup>C/<sup>12</sup>C ratio 142 143 in CO<sub>2</sub>. The sample was injected into a Trace-GC with a Precon Interface connected with a continuousflow isotope-ratio mass spectrometer (IRMS; Thermo Finnigan DELTA XP<sup>Plus</sup>, Bremen, Germany). 144 Separation of CO<sub>2</sub> from N<sub>2</sub>O and other gases was performed with a Pora PLOT Q column (27.5 m 145 146 length; 0.32 mm i.d.; Varian) at 25 °C using He as carrier gas. Laboratory standards were prepared using CO<sub>2</sub> gas diluted in N<sub>2</sub> and calibrated against the international IAEA CH-6 reference material 147 (International Atomic Energy Agency, Vienna, Austria) via elemental analysis IRMS (EA-IRMS). The 148 remaining gas in the original 12-ml vials was then analyzed for  ${}^{13}C/{}^{12}C$  ratio of CH<sub>4</sub>. The same protocol 149 150 was used as described above with the exception that an additional, manual cold trap was cooled by liquid nitrogen in the Precon unit to purge the sample gas of  $O_2$  and  $N_2$  while trapping CH<sub>4</sub> in the loop. 151 Additionally,  $CH_4$  was oxidized to  $CO_2$  by reaction with nickel oxide at 1000°C. The precision (one 152 153 standard deviation) of standard gas for 10 consecutive measurements is ~0.1‰. Isotope ratios are reported in terms of <sup>13</sup>C at% values, where at%  $^{13}C = {}^{13}C/({}^{12}C+{}^{13}C)*100$ . Natural abundance of  ${}^{13}C$  was 154 taken into account by analysis of [<sup>12</sup>C]cellobiose-incubated samples as controls. <sup>13</sup>C at% values were 155 used to calculate the production of <sup>13</sup>C-CH<sub>4</sub> and <sup>13</sup>C-CO<sub>2</sub> based on the overall production of CH<sub>4</sub> and 156 157 CO<sub>2</sub> (Table S1).

158

# 159 RNA extraction

160 The method of RNA extraction was modified from previously published methods (37-39). Frozen peat
161 (0.5 g wet weight) was added to two 2-ml tubes containing 0.5 cm of quartz sand. Both tubes received

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|     | ······································   |
|-----|--|
| 163 | [PVP], 100 mM Tris-HCl [pH 8.0], 25 mM EDTA, 2.0 M NaCl, 0.5 g l $^{-1}$ spermidine, 2% $\beta$ -                      |
| 164 | mercaptoethanol) (40) and 350 µl of phenol (pH 8). Cell lysis was carried out in FastPrep (Qbiogene,                   |
| 165 | Illkirch, France) with setting 5.5 m s <sup>-1</sup> for 30 s. After centrifugation at 20 800 g at 4 °C for 5 min, the |
| 166 | supernatants from both tubes were combined and 700 $\mu$ l of phenol-chloroform-isoamylalcohol                         |
| 167 | (50:49:1) was added. Samples were mixed by inverting the tube repeatedly and centrifuged as above                      |
| 168 | for 3 min. Chloroform-isoamyl alcohol (700 $\mu$ l, 24:1) was added to the supernatant. After mixing and               |
| 169 | centrifugation, the supernatant was passed through a polyvinyl polypyrrolidone (PVPP) column by                        |
| 170 | centrifugation at 1300 g for 3 min. The flow-through was precipitated with 1 volume of 20%                             |
| 171 | polyethylene glycol in 2.5 M NaCl and 20 $\mu$ g of glycogen (RNA grade, ThermoScientific) on ice for                  |
| 172 | 1.5 h. After centrifugation at 20 800 g at 4 °C for 30 min, the pellet was washed with 70% ethanol and                 |
| 173 | dissolved in 50 $\mu$ l of diethylpyrocarbonate (DEPC)-treated H <sub>2</sub> O. Co-extracted DNA was removed by       |
| 174 | treating RNA with DNase I (Thermo Scientific). Concentration of RNA was determined with Qubit                          |
| 175 | RNA HS Assay Kit and Qubit fluorometer (Invitrogen).   |
|     |  |

350 µl of lysis buffer (2% cetyl trimethylammonium bromide [CTAB], 2% polyvinyl pyrrolidone

176

# 177 Density gradient centrifugation of RNA

Separation of <sup>13</sup>C-labeled RNA from <sup>12</sup>C-RNA was carried out by density gradient centrifugation in cesium trifluoroacetate (CsTFA) gradients (41) for RNA from day 7 and day 14. Gradient medium was prepared by mixing 4.8 ml of CsTFA (2.0 mg ml<sup>-1</sup>, GE Healthcare), 3.4% formamide, and 938 µl of gradient buffer (0.1 M Tris-HCl, 0.1 M KCl, 1 mM EDTA) for a final volume of 6 ml. Density of CsTFA and the final gradient medium were checked before use by measuring the refractory index. Gradient medium and RNA sample (500 ng RNA in a volume of 54.5 µl) were loaded into 6-ml

184 polyallomer crimp tubes (Sorvall) and centrifuged at 129 000 g for 65 h in a Discovery 100 centrifuge

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| 185 | with a TV-1665 rotor (Sorvall). Each centrifugation included two control tubes without RNA or with                            |
|-----|---|
| 186 | RNA from a cultivated strain for determining the density of fractions. After centrifugation, gradients                        |
| 187 | were fractionated into 14 fractions of 400 $\mu$ l by pumping water to the top of the tube using a Gilson                     |
| 188 | Minipuls 3 (speed 3, ca. 0.5 ml min <sup>-1</sup> ) and collecting the fractions from the bottom. Densities of the            |
| 189 | fractions from the control tubes were measured by a refractometer and by weighing the fractions.                              |
| 190 | RNA in the fractions was precipitated with isopropanol and 20 $\mu g$ of glycogen at -20 $^{\circ}\mathrm{C}$                 |
| 191 | overnight, centrifuged at 20 800 g at 4 °C for 30 min, washed with 70% ethanol, and dissolved in 15 $\mu l$                   |
| 192 | of DEPC-H <sub>2</sub> O. Concentration of RNA was determined with Qubit RNA HS Assay Kit and Qubit                           |
| 193 | fluorometer (Invitrogen). Fractions containing <sup>13</sup> C-labeled RNA and <sup>12</sup> C- RNA were identified by        |
| 194 | comparing RNA concentrations in the density profiles of [ <sup>13</sup> C] and [ <sup>12</sup> C]cellobiose-incubated samples |
| 195 | (Fig. S1). Two to three fractions were pooled to form the final heavy ( $^{13}$ C-enriched) and light ( $^{12}$ C)            |
| 196 | RNA fractions of each sample.   |
| 197 |   |
| 198 | Reverse transcription and 16S rRNA gene amplicon sequencing   |
| 199 | The heavy and light RNA fractions and RNA from initial peat (4 $\mu$ l aliquots) were reverse transcribed                     |

with Maxima H Minus reverse transcriptase (200 U, Thermo Scientific) and random hexamers (Thermo
Scientific) following the manufacturer's instructions.

- 202 Bacterial 16S rRNA gene PCR for reverse transcription products was carried out as a two-step PCR,
- 203 where the first step primers (341f and 805r) (42) contained adapters for introducing Illumina adapters
- and dual barcodes in the second step. The first step PCR primers were thus adapter-341f (5'-

# 205 ACACTCTTTCCCTACACGACGCTCTTCCGATCTNNNNCCTACGGGNGGCWGCAG-3';

adapter sequence followed by primer sequence in bold) and adapter-805r (5'

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| 207 | AGACGTGTGCTCTTCCGATCTGACTACHVGGGTATCTAATCC-3'). The first step PCR was                                     |
|-----|--|
| 208 | carried out in duplicate in 20-µl reactions containing 1 $\times$ Q5 reaction buffer, 0.2 mM dNTPs, 0.5 µM |
| 209 | primers, 0.4 U Q5 high-fidelity DNA polymerase (New England Biolabs) and 1 $\mu$ l reverse transcription   |
| 210 | product (diluted if necessary) as template. Cycling conditions were 98 °C for 1 min, followed by 15-20     |
| 211 | cycles of 98 °C for 10 s, 62 °C for 30 s, 72 °C for 30 s, and a final extension at 72 °C for 2 min. The    |
| 212 | duplicate products were pooled and purified with Agencourt AMPure XP purification system                   |
| 213 | (Beckman Coulter). The second PCR step with barcoded primers (forward                                      |
| 214 | AATGATACGGCGACCACCGAGATCTACAC-[index]-ACACTCTTTCCCTACACGACG; reverse                                       |
| 215 | CAAGCAGAAGACGGCATACGAGAT-[index]-  |
| 216 | GTGACTGGAGTTCAGACGTGTGCTCTTCCGATCT), which bind to the first-step adapters and                             |
| 217 | incorporate Illumina adapters, was carried out using 1 µl of the first product (dilution 1:1-1:50          |
| 218 | depending on concentration) as template with 0.1 $\mu M$ of primers, annealing temperature of 66 °C, and   |
| 219 | 12-17 cycles. Products were purified as above and quantified with PicoGreen kit (Life Technologies).       |
| 220 | Products were sequenced at the SciLifeLab SNP/ SEQ sequencing facility at Uppsala University,              |
| 221 | Uppsala, Sweden on Illumina MiSeq (2×300 bp).  |
| 222 |  |
| 223 | Sequence analysis  |
| 224 | After raw amplicon sequencing data was demultiplexed into individual samples, sequence pairs were          |
| 225 | assembled using an in-house bioinformatic pipeline (43). The pipeline further removed sequences with       |

226 missing primers and unassigned base pairs. The quality filtering removed on 17% of reads, leaving 2.16

227 million reads of 400-430 bp and on average 41379 reads per sample. The quality filtered assembled

- 228 reads were clustered into operational taxonomical units (OTUs) and filtered for chimeras using
- 229 UPARSE (1.5 % sequence dissimilarity cutoff) (44), which resulted in 9243 OTUs. Singletons were

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removed. Taxonomy was assigned using CREST (45) and the ribosomal sequence database SilvaMod based on the release 106 of the SILVA non-redundant SSURef database (46). The sequence data were sub-sampled to 13226 reads per sample, the number of reads in the sample with the fewest reads after all the filtering steps, using the command sub.sample in mothur (v. 1.35.1)(47).

234

# 235 Statistical analyses

We used t-tests to compare  $CH_4$  and  $CO_2$  production rate of the original peat of the oligotrophic and mesotrophic fen (n=3). Methane production was  $log_{10}$  transformed before testing. <sup>13</sup>C-CH<sub>4</sub> and <sup>13</sup>C-CO<sub>2</sub> production rates between peat types during weeks 1, 2, and 3 (n=3) were tested with repeated measures analysis of variance using function lme in R package nlme (48) with week as a random factor. All analyses were carried out in R v.3.2.0 (49). P values were considered statistically significant at p < 0.05.

Bacterial community in <sup>13</sup>C and <sup>12</sup>C fractions at days 7 and 14 and in initial peat of the fens were visualized by non-metric multidimensional scaling (NMDS) plots based on Bray-Curtis distances of OTU data subsampled to 13226 reads per sample (n=3). Differences with peat type, fraction, and time point were tested with permutational multivariate analysis of variance (PERMANOVA, function adonis in R, Bray-Curtis distances). Analyses were carried out in R with vegan package (v. 2.0) (50).

The OTUs enriched in the <sup>13</sup>C fractions compared to the <sup>12</sup>C fractions were identified with R package edgeR (51). Data were not subsampled for these analyses (52). Filtering was applied to keep only OTUs with more than five reads in at least four samples within each comparison of 12 samples, i.e., within <sup>13</sup>C and <sup>12</sup>C fractions of two time points with three replicates. The filtered data was

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| 252 | normalized by the relative log expression (RLE) method as implemented in standard edgeR protocols.                                  |
|-----|---|
| 253 | We used the general linear model (glm) approach to account for the two time points of 7 and 14 days.                                |
| 254 | Trended dispersion was estimated with the power method and the tagwise dispersion with df=5. In                                     |
| 255 | testing for enriched ('differentially expressed') OTUs, we accepted OTUs with $p < 0.01$ , false                                    |
| 256 | discovery rate (FDR) $\leq$ 0.05 and log <sub>2</sub> fold change (logFC) $\geq$ 2 as enriched in the <sup>13</sup> C fractions. To |
| 257 | exclude OTUs migrating to the <sup>13</sup> C fraction without actual <sup>13</sup> C incorporation, we tested for OTUs             |
| 258 | enriched in the fractions corresponding to the density of the <sup>13</sup> C fraction in samples incubated with                    |
| 259 | $[^{12}C]$ cellobiose. Based on this test, one OTU from the mesotrophic peat was removed from the set of                            |
| 260 | enriched OTUs. In further analyses, the read numbers of the enriched OTUs are from the subsampled                                   |
| 261 | data. Sequences similar to representative sequences of enriched OTUs were searched using Blast.                                     |
| 262 |   |

### 263 Accession number

264 Raw amplicon sequences were deposited in the GenBank sequence read archive under accession 265 number SRP075161.

266

Results 267

### Methane and anaerobic CO<sub>2</sub> production 268

269 Potential CH<sub>4</sub> production in initial measurements before cellobiose addition was higher in peat from the

- 270 20-30 cm layer than in the 10-20 cm layer (data not shown). The depth of 20-30 cm was therefore
- 271 chosen for the labeling experiment to represent the methanogenic peat layer. Before cellobiose
- 272 amendment, the mesotrophic fen showed higher  $CH_4$  production (36 ± 18 nmol grams dry weight

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| 273 | $(gdw)^{-1} h^{-1}$ , mean $\pm$ SD) than the oligotrophic fen (1.2 $\pm$ 0.8 nmol gdw <sup>-1</sup> h <sup>-1</sup> ; t-test p = 0.003). The rates    |
|-----|--|
| 274 | of anaerobic CO <sub>2</sub> production did not significantly differ between mesotrophic $(116 \pm 72 \text{ nmol gdw}^{-1} \text{ h}^{-1})$           |
| 27: | <sup>1</sup> ) and oligotrophic peat ( $61 \pm 10 \text{ nmol gdw}^{-1} \text{ h}^{-1}$ , t-test p = 0.32). Rates of CO <sub>2</sub> production of the |
| 270 | original unamended peat were on average 3 times higher than CH <sub>4</sub> production in the mesotrophic fen  |
| 277 | and 50 times higher in the oligotrophic fen.   |
| 278 | After the addition of [ <sup>13</sup> C]cellobiose, <sup>13</sup> C-enrichment of CH <sub>4</sub> demonstrated consumption                             |
| 279 | of cellobiose with methanogenesis as a terminal process (Fig. 1A). The $^{13}$ C-enrichment of CH <sub>4</sub>   |
|     |  |

increased from day 7 to day 28, whereas the enrichment of CO<sub>2</sub> remained fairly stable over time. The 280 level of enrichment did not vary with peat type (Fig. 1A). The mesotrophic peat had consistently higher 281  $^{13}$ C-CH<sub>4</sub> production, i.e., CH<sub>4</sub> originating from labeled cellobiose, than the oligotrophic peat (p = 282 283 0.0001, Fig. 1B). Methane production was highest in the first week and then decreased in both peat 284 types. Anaerobic  ${}^{13}C-CO_2$  production did not differ with peat type (p = 0.84, Fig. 1B). In both peat types, CO<sub>2</sub> production increased after the first week and was highest in the second week. Accordingly, 285 in the first week the ratio of  ${}^{13}$ C-CO<sub>2</sub> and  ${}^{13}$ C-CH<sub>4</sub> production was  $1.4 \pm 0.3$  in the mesotrophic peat 286 and  $5 \pm 2$  in the oligotrophic peat, but in the second week the ratio increased to  $7 \pm 5$  in the 287 288 mesotrophic peat and  $32 \pm 24$  in the oligotrophic peat. At weeks 3 and 4 the ratio decreased to  $14 \pm 6$  in the oligotrophic peat and  $5 \pm 2$  in the mesotrophic peat, still being higher than in the first week. 289

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### 291 **Bacterial communities**

| 292 | The bacterial community of the original peat differed between the oligotrophic and mesotrophic fen           |
|-----|--|
| 293 | (Fig. 2; PERMANOVA $R^2 = 0.48$ , p = 0.001). The most common bacterial phyla in the oligotrophic fen        |
| 294 | were Parcubacteria (Candidate division OD1, $15.4 \pm 2.2\%$ of reads), Planctomycetes ( $13.2 \pm 3.6\%$ ), |

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Acidobacteria  $(10.8 \pm 4.2\%)$ .

Light fractions at day 7 differed from the original peat (PERMANOVA  $R^2 = 0.17$ , p = 0.007). No 301 change was observed between day 7 and day 14 in the overall community (PERMANOVA  $R^2 = 0.01$ , p 302 = 0.80) or in the mesotrophic peat despite the clear decrease in  $CH_4$  production (PERMANOVA  $R^2$  = 303 0.06, p = 0.35). Unlabeled controls that received [<sup>12</sup>C]cellobiose, used to identify the density gradient 304 fractions containing <sup>13</sup>C-RNA and to exclude OTUs migrating to the heavy fraction without <sup>13</sup>C 305 306 incorporation, grouped with the light fractions (data not shown). 307 Looking at the fold changes of the OTU abundances in the heavy and light fractions 308 revealed that *Firmicutes* and *Betaproteobacteria* were strongly represented in the heavy fractions (Fig. 309 3, Table S3). Several groups common in the original peat were primarily detected in the light fractions 310 (Planctomycetes, Deltaproteobacteria, Chloroflexi) or were very rare in incubated peat 311 (Parcubacteria/Candidate division OD1). To identify the most likely cellobiose-C consumers, we 312 looked in more detail for those OTUs consistently more abundant in the heavy fractions compared to 313 the light fractions by using an approach developed for detecting differential gene expression. In the 314 mesotrophic peat, we identified 70<sup>13</sup>C-enriched OTUs, 44 of which belonged to *Firmicutes* (Table 1, Table S4, Table S5). The oligotrophic peat revealed higher number and higher taxonomic variety of 315 <sup>13</sup>C-enriched OTUs than the mesotrophic peat with 118 enriched OTUs and Acidobacteria, 316

and Acidobacteria  $(10.2 \pm 2.0\%)$  (Table S2). In the mesotrophic fen, the groups with the highest

and light (<sup>12</sup>C) fractions, OTUs differed between the fractions (PERMANOVA  $R^2 = 0.34$ , p = 0.001)

and between the oligotrophic and mesotrophic peat (PERMANOVA  $R^2 = 0.27$ , p = 0.001) (Fig. 2).

After incubation with [<sup>13</sup>C]cellobiose and separation of RNA into heavy (<sup>13</sup>C-enriched)

relative abundance were Deltaproteobacteria  $(17.6 \pm 5.9\%)$ , Chloroflexi  $(10.9 \pm 4.6\%)$ , and

317 Verrucomicrobia, Firmicutes, Betaproteobacteria, and Alphaproteobacteria as the largest groups.

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318 Eleven OTUs in the oligotrophic peat and six OTUs in the mesotrophic peat became more enriched from day 7 to day 14 (Table 1, Table S5), further supporting their <sup>13</sup>C-labeling. 319 320 Firmicutes classified to the families Clostridiaceae and Veillonellaceae were identified as 321 prominent cellobiose-C consumers in both peat types. In unincubated peat, *Firmicutes* formed only 0.1-322 1.0% of the total community, but by day 7 they had increased to 21-69% in the heavy fractions (Table S3). Clostridiaceae in the oligotrophic peat, including the two most abundant <sup>13</sup>C-enriched OTUs 323 324 (Otu2, Otu4), had high sequence similarities to several acid-tolerant and sugar-fermenting Clostridium species (Table S4, S5). In the mesotrophic peat, the two most abundant *Clostridium* OTUs (Otu1, 325 326 Otu3) showed similarities of 99-100% to uncultured clostridia from wetland soils (26, 28, 53, 54). Both 327 oligotrophic and mesotrophic peat showed Veillonellaceae OTUs similar to genera Pelosinus, 328 Propionispira and Psychrosinus (Table S4), but several additional Veillonellaceae OTUs from the 329 mesotrophic peat had no close matches to described species. Only three enriched Firmicutes OTUs 330 were unique to the oligotrophic peat, and they were similar to "Psychrosinus fermentas" (Otu110, 331 1663, 5263; Table S4, S5). The mesotrophic peat, on the other hand, showed 24 unique enriched 332 Clostridiaceae and Veillonellaceae OTUs and two additional families: Lachnospiraceae and 333 Ruminococcaceae. All four Ruminococcaceae OTUs (Otu40, 64, 145, 316) were 94-96% similar to 334 Ethanoligenens harbinense and 96-98% similar to a sequence from acidic peat (HG324862, 28). The 335 closest matches to Lachnospiraceae (Otu119) were gut microbes. The majority of the <sup>13</sup>C-enriched *Betaproteobacteria* OTUs in both peat types were 336 337 classified as Neisseriales and Rhodocyclales. Many of these OTUs (Table S4) showed high sequence 338 similarity to two wetland isolates: fermentative Paludibacterium yongneupense growing at lower pH

than related strains (55) and aerobic Uliginosibacterium gangwonense (56).

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340 Enriched Alphaproteobacteria OTUs belonged to Rhodospirillales and Rhizobiales. Both 341 the oligotrophic and the mesotrophic peat revealed several OTUs (Table S4) highly similar to 342 Telmatospirillum siberiense, an acidotolerant fermentative peat isolate (57). Rhizobiales OTUs were mostly detected as <sup>13</sup>C-enriched in the oligotrophic peat, and most of them were similar to one of two 343 344 stalked fermentative strains able to use Fe(III) as an electron acceptor: Rhizomicrobium electricum (58; 345 Otu761, 463, 1222) and Rhizomicrobium palustre (59; Otu32).

Only two <sup>13</sup>C-enriched Acidobacteria OTUs, classified to Holophagae, were detected in 346 the mesotrophic peat (Otu63, Otu2366). In the oligotrophic peat, we identified 26 <sup>13</sup>C-enriched 347 348 Acidobacteria OTUs, and 12 of them showed 97-99% sequence similarity to cellulolytic Telmatobacter 349 bradus isolated from a bog with pH 4 and fermenting sugars including cellobiose (60) (Table S4). The 350 rest of the Acidobacteria OTUs belonging to classes Acidobacteria and Holophagae were highly 351 similar (99-100%) to sequences from acidic peat (26, 28, 33, 61, 62), alpine tundra soil (63), or paddy 352 soil (Table S5).

353 Verrucomicrobia were identified as consumers of cellobiose-derived carbon only in the 354 oligotrophic fen. Fourteen out of twenty OTUs were classified to subphylum 3 (OPB35 soil group) 355 (Table S5). These OTUs showed only  $\leq 91\%$  sequence similarity to the described members of 356 subplyum 3, which all are aerobic (64-66). In addition, OTUs identified as Spartobacteria (Otu470, 357 Otu1065) and Opitutae (Otu224, 266, 364, 613) were detected, both of which include a species 358 fermenting sugars and polysaccharides (67, 68). All Verrucomicrobia OTUs were most similar to 359 uncultured Verrucomicrobia from various soil, wetland and freshwater environments.

360 Bacteroidetes, previously identified as a major fermentative group in peat (28), were not prominent among our <sup>13</sup>C-enriched OTUs. However, two relatively minor *Bacteroidetes* OTUs (Otu24, 361

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Otu198) in the mesotrophic peat identified as *Porphyromonadaceae* became more <sup>13</sup>C-enriched with
time (Table 1, Table S5).

364

# 365 Discussion

366 We compared consumers of cellobiose-derived carbon in two peat types with similar anaerobic  $CO_2$ 367 production but distinct rates of CH<sub>4</sub> production, suggesting differences in the processes of organic 368 matter degradation. The ratio of CO<sub>2</sub> to CH<sub>4</sub> production of close to 1 in the mesotrophic peat during the 369 first week of the incubation implies that methanogenesis was the dominant terminal process. The 370 considerably higher CO<sub>2</sub>/CH<sub>4</sub> ratios in the oligotrophic peat throughout the incubation, ranging from 2 371 to 59, indicate that carbon was also directed to non-methanogenic processes. Our CO<sub>2</sub>/CH<sub>4</sub> ratios of 372 both original and cellobiose-amended peat were higher than those of Galand et al. (21) for the same 373 sites but support the same pattern of lower contribution of methanogenesis as a terminal process in the 374 oligotrophic peat.

375 A clear separation of the bacterial community composition with peat type persisted with 376 cellobiose amendment, and the results showed distinct cellobiose-C consuming taxa with peat type and 377 a higher variety of taxa in the oligotrophic peat. These included Acidobacteria and Verrucomicrobia, which were present in the mesotrophic peat at similar or higher relative abundances as in the 378 379 oligotrophic peat but were not as heavily labeled (Table S2, Table S3), suggesting they did not 380 assimilate cellobiose-derived carbon to the same extent as in the oligotrophic peat. Verrucomicrobia 381 have been associated with plant polymer degradation in soils including anoxic peat (32, 69) and in the 382 case of Opitutaceae with glucose fermentation in peat (33). Subphylum 3, the most abundant verrucomicrobial group detected here as cellobiose <sup>13</sup>C-consumers, occurs in soil and peat (70-72) but 383

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contains no previously reported fermenters. According to recent metagenomic data, some members of
this group possess genes for polysaccharide hydrolysis (73). Our results show that subphylum 3
contains members able to assimilate cellobiose carbon under anoxic, oligotrophic conditions, extending
the physiologies for subphylum 3 and revealing potential novel peatland fermenters.

388 Acidobacteria are common in anoxic peat and more prominent with lower pH and oligotrophy (4, 33, 34, 74-77). Accordingly, they were more prominent in processing cellobiose-C in 389 390 the oligotrophic peat than in the mesotrophic peat. Acidobacteria closely related to Telmatobacter bradus, detected as <sup>13</sup>C-enriched exclusively in the oligotrophic peat, are emerging as important 391 anaerobic and aerobic degraders of cellulose-derived carbon in peatlands with pH < 5 (28, 78). 392 393 Additionally, other Acidobacteriaceae and Holophagaceae appeared to participate in anaerobic carbon 394 processing, but the roles of these taxa are more difficult to determine. Known Holophagaceae have diverse physiologies, but members of this group also assimilated <sup>13</sup>C from cellulose in acidic peat (28). 395 Firmicutes assimilating <sup>13</sup>C were particularly prominent and varied in the mesotrophic 396

397 peat, and those found in the oligotrophic peat were mostly a subset of this larger variety. Nutrient status 398 has been previously shown to affect the distribution of *Clostridia* in a freshwater marsh between eutrophic and oligotrophic soils (53). The two *Firmicutes* families detected as <sup>13</sup>C consumers only in 399 400 the mesotrophic peat, Ruminococcacae and Lachnospiraceae, are best known to inhabit digestive tracts 401 of mammals (79). Interestingly, the occurrence of *Ruminococcaceae* and *Lachnospiraceae* in sheep 402 rumen was associated with high CH<sub>4</sub> emission (80), fitting their detection in the mesotrophic peat with 403 higher CH<sub>4</sub> production. Ruminococcaceae have been indicated as fermenters in peat more acidic than 404 our sites (28, 33), suggesting that the lack of detection in the oligotrophic peat may not be related to the 405 slightly lower pH. Lachnospiraceae have been detected in degradation of rice straw (81) but not

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Our <sup>13</sup>C-enriched Firmicutes OTUs closely matched sequences from other peatlands and 408 409 acid-tolerant wetland isolates. These included *Clostridiaceae*, which are well-known saccharolytic and 410 cellulolytic fermenters and have been suggested to represent important fermenters in acidic peat (26, 28, 33, 52). Another major group of Firmicutes were Veillonellaceae, which have Gram-negative cell 411 412 walls and have so far been rarely detected in peat or soil habitats, although they have been actively involved in rice straw degradation (81) and detected as minor peat taxa (33, 83). Veillonellaceae may 413 414 play a role as propionate-producing fermenters in peat: Several of our OTUs were similar to 415 propionate-producing genera Pelosinus, Psychrosinus, and Propionispira (Table S4), and 416 Propionispira-related glucose fermenters were previously detected in acidic peat under propionate-417 producing conditions (26). Veillonellaceae also include homoacetogens producing acetate from H<sub>2</sub> and CO<sub>2</sub>, but none of our OTUs were similar to known acetogenic strains. 418 419 The large amount of <sup>13</sup>C-CO<sub>2</sub> detected throughout the incubation in the oligotrophic peat could be the result of incomplete degradation or degraders using humic substances (HS) or inorganic 420 421 compounds as electron acceptors. Incomplete degradation, where electron donors and acceptors are 422 organic substances and which results in build-up of fermentation products instead of CH<sub>4</sub> production, is 423 one of the main proposed sources for the excess  $CO_2$  production in peat (13, 21). The various potentially fermentative groups that assimilated <sup>13</sup>C support incomplete degradation as a relevant 424 425 source. The role of HS reduction (84) is more difficult to evaluate, as HS reducers are poorly known 426 and include Fe(III)-reducing, sulfate reducing and fermentative bacteria (85-88). The labeled taxa in 427 the oligotrophic peat included Holophagaceae and Anaeromyxobacter (Deltaproteobacteria) with 428 members able to reduce HS analogue anthraquinone-2,6-disulfonate (AQDS; 89). None of the labeled

430

| Accepted                                 | 432 | reducers could also be speculated to reduce HS, although most have not been tested for it, and a   |
|--|-----|--|
| Scep                                     | 433 | 3 acidobacterium Paludibaculum fermentas does not reduce AQDS (94). Whether the groups dis   |
| Ă  | 434 | above are involved in HS or Fe(III) reduction and resposible for the excess CO <sub>2</sub> production shou  |
|  | 435 | addressed in future studies. Other <sup>13</sup> C-CO <sub>2</sub> sources that cannot be ruled out are <sup>13</sup> C-CO <sub>2</sub> production |
|  | 436 | microbes not assimilating <sup>13</sup> C, production by organisms other than bacteria such as non-methano   |
|  | 437 | archaea, or anaerobic CH <sub>4</sub> oxidation (95). Despite attempts, anaerobic CH <sub>4</sub> oxidation has not bee                            |
|  | 438 | conclusively verified in this site (K. Peltoniemi, personal communication).  |
| Appred and crivironmenia<br>Microbiology | 439 | In the mesotrophic peat, CH <sub>4</sub> production decreased and CO <sub>2</sub> production increased   |
|  | 440 | drastically after the first week, but we observed no shift in the bacterial community. The addition  |
|  | 441 | CO <sub>2</sub> was therefore possibly produced by <i>Firmicutes</i> and <i>Proteobacteria</i> that assimilated <sup>13</sup> C und                |
| Applied                                  | 442 | methanogenic conditions and remained active under lower CH4 production. The decrease in CH4  |
| •  | 443 | production may be due to inhibition of methanogenic activity by accumulation of fermentation   |
|  | 444 | products such as organic acids, as has been observed in other peat incubations (28, 96). The incu  |
|  | 445 | mesotrophic peat revealead notable amounts of deltaproteobacterial Fe(III) and sulfate reducers,   |

429 taxa resembled known sulfate reducers but instead contained several potential Fe(III) reducers:

431 92). Fermentative Fe(III) reducers have been shown to occur in an acidic fen (93). The putative Fe(III) ice HS, although most have not been tested for it, and a group tas does not reduce AQDS (94). Whether the groups discussed ction and resposible for the excess CO<sub>2</sub> production should be  $D_2$  sources that cannot be ruled out are  ${}^{13}C-CO_2$  production by n by organisms other than bacteria such as non-methanogenic . Despite attempts, anaerobic CH<sub>4</sub> oxidation has not been oniemi, personal communication).

Rhizomicrobium, Clostridia related to C. saccharobutylicum, and Acidobacteria of groups 1 and 3 (90-

served no shift in the bacterial community. The additional <sup>13</sup>C-Firmicutes and Proteobacteria that assimilated <sup>13</sup>C under ctive under lower CH<sub>4</sub> production. The decrease in CH<sub>4</sub> ethanogenic activity by accumulation of fermentation en observed in other peat incubations (28, 96). The incubated unts of deltaproteobacterial Fe(III) and sulfate reducers, such 446 as Geobacteraceae (2-4% vs. < 1% in oligotrophic peat), Syntrophaceae (3-5% vs. 1-3%), Syntrophobacteraceae (0.9-1.3% vs. 0.4-0.7%), and Desulfobacteraceae (0.1-0.5% vs. < 0.1%). These 447 taxa were, however, not enriched in the <sup>13</sup>C-fractions and thus were most likely not producing the 448 additional  ${}^{13}$ C-CO<sub>2</sub> from  $[{}^{13}$ C]cellobiose, but they may be contributing to CO<sub>2</sub> production overall from 449 450 endogenous substrates.

To avoid known sources of bias in SIP experiments, we added the <sup>13</sup>C-substrate as 451 452 repeated smaller pulses instead of a large single pulse and used RNA-SIP instead of DNA-SIP to allow shorter incubation times. To exclude OTUs enriched in the heavy fraction for reasons other than <sup>13</sup>C-453 454 labeling such as high GC percentage, we sequenced the fractions corresponding to the heavy fractions 455 in the unlabeled samples and removed such OTUs. Another source of bias, cross-feeding, should have 456 been reduced as the incubation flasks were flushed during the weekly sampling. Despite removal of the 457 gases, cross-feeding could still take place via organic acids. Therefore the detected labeled taxa may 458 include both primary fermenters, using cellobiose and glucose, and secondary fermenters consuming 459 the products of primary fermenters such as organic acids. Because cellobiose-C consumers were 460 identified based on enrichment in the heavy fractions compared to the light fractions, our analysis focuses on the most strongly <sup>13</sup>C-labeled taxa. Therefore the analysis would miss OTUs assimilating 461 both <sup>13</sup>C-carbon and endogenous unlabeled carbon, thus appearing in the heavy fractions but in 462 463 addition strongly or predominantly in the light fractions.

Several potentially fermentative groups abundant in the original peat were not labeled 464 with <sup>13</sup>C from cellobiose. *Planctomycetes*, largely related to *Isosphaera*, became the most abundant 465 group in oligotrophic peat with cellobiose amendment, but only one rare OTU was identified as <sup>13</sup>C-466 enriched. Planctomycetes have been described as aerobic, rarely fermentative carbohydrate degraders 467 468 in soil and peat (73, 97, 98) and proposed to be glucose fermenters in peat (33). Our results match those of Schmidt et al. (28), where *Planctomycetes* were not labeled from <sup>13</sup>C-cellulose. Similarly, 469 470 Chloroflexi classified as Anaerolineae and Caldilineae, which consist of filamentous anaerobes that 471 ferment sugars and were enriched with glucose in peat (33, 99), were mostly detected in the light 472 fractions of the mesotrophic peat. Strong occurrence in the light fractions suggests that Planctomycetes 473 and Chloroflexi were active but largely not assimilating cellobiose-derived carbon. Parcubacteria

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474 (Candidate division OD1), which have been proposed to be fermentative (100, 101), were abundant in
475 the original oligotrophic peat but were rare and unlabeled after the incubation. Members of this group
476 were suggested to be symbiotic (102), which could explain why they thrive in natural but not in
477 incubated peat.

478 To conclude, we showed that anaerobic degradation in peat, exemplified by cellobiose-C 479 consumers, can involve a clearly distinct set of bacteria depending on the amount of CH<sub>4</sub> produced. 480 With higher CH<sub>4</sub> production in the mesotrophic peat, cellobiose-derived carbon was mainly processed 481 by well-known fermenters within *Firmicutes* and by *Proteobacteria*. When CO<sub>2</sub> production by far 482 exceeded CH<sub>4</sub> production in the oligotrophic peat, more unconventional degraders or fermenters such 483 as Telmatobacter-related Acidobacteria and subphylum 3 of Verrucomicrobia were prevalent. The groups we detected as <sup>13</sup>C-labeled in the oligotrophic peat, particularly those with known potential for 484 485 Fe(III) or HS reduction, such as Acidobacteria, Holophagaceae, Deltaproteobacteria, Rhizobiales, and 486 Clostridiaceae, should be further addressed in future studies to resolve the processes and electron 487 acceptors behind the anaerobic CO<sub>2</sub> production in acidic peatlands. By differentiating anaerobic 488 bacteria active in cellobiose processing from the rest of the community, this study contributes to 489 assigning functions to uncharacterized bacteria in peat and to understanding the microbiological basis 490 of differing levels of CH<sub>4</sub> production in peatlands.

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500 References

501 1. Kirschke S, Bousquet P, Ciais P, Saunois M, Canadell JG, Dlugokencky EJ, Bergamaschi P,

Bergmann D, Blake DR, Bruhwiler L. 2013. Three decades of global methane sources and sinks. Nat
Geosci 6:813-823.

504 2. Turetsky MR, Kotowska A, Bubier J, Dise NB, Crill P, Hornibrook ER, Minkkinen K, Moore

505 TR, Myers-Smith IH, Nykänen H, Olefeldt D, Rinne J, Saarnio S, Shurpali N, Tuittila E-S,

506 Waddington JM, White JR, Wickland KP, Wilmking M. 2014. A synthesis of methane emissions

from 71 northern, temperate, and subtropical wetlands. Global Change Biol **20**:2183-2197.

508 3. Whalen SC. 2005. Biogeochemistry of methane exchange between natural wetlands and the

509 atmosphere. Environ Eng Sci 22:73-94.

510 4. Urbanová Z, Bárta J. 2016. Effects of long-term drainage on microbial community composition

511 vary between peatland types. Soil Biol Biochem 92:16-26.

5. Valentine DW, Holland EA, Schimel DS. 1994. Ecosystem and physiological controls over

513 methane production in northern wetlands. J Geophys Res 99:1563-1571.

514 6. Godin A, McLaughlin JW, Webster KL, Packalen M, Basiliko N. 2012. Methane and

515 methanogen community dynamics across a boreal peatland nutrient gradient. Soil Biol Biochem **48:**96-

516 105.

Applied and Environmental Microbiology

AEM

| 517 | 7. Yrjälä K, Tuomivirta TT, Juottonen H, Putkinen A, Lappi K, Tuittila E-S, Penttilä T,                  |
|-----|--|
| 518 | Minkkinen K, Laine J, Peltoniemi K, Fritze H. 2011. CH4 production and oxidation processes in a          |
| 519 | boreal fen ecosystem after long-term water table drawdown. Global Change Biol 17:1311-1320.              |
| 520 | 8. Galand PE, Fritze H, Conrad R, Yrjälä K. 2005. Pathways for methanogenesis and diversity of           |
| 521 | methanogenic archaea in three boreal peatland ecosystems. Appl Environ Microbiol 71:2195-2198.           |
| 522 | 9. Yavitt JB, Yashiro E, Cadillo-Quiroz H, Zinder SH. 2012. Methanogen diversity and community           |
| 523 | composition in peatlands of the central to northern Appalachian Mountain region, North America.          |
| 524 | Biogeochemistry 109:117-131.   |
| 525 | 10. Kettunen A, Kaitala V, Lehtinen A, Lohila A, Alm J, Silvola J, Martikainen PJ. 1999.                 |
| 526 | Methane production and oxidation potentials in relation to water table fluctuations in two boreal mires. |
| 527 | Soil Biol Biochem <b>31:</b> 1741-1749.  |
| 528 | 11. Rooney-Varga JN, Giewat MW, Duddleston KN, Chanton JP, Hines ME. 2007. Links between                 |
| 529 | archaeal community structure, vegetation type and methanogenic pathway in Alaskan peatlands. FEMS        |
| 530 | Microbiol Ecol <b>60:</b> 240-251.   |
| 531 | 12. Laine J, Minkkinen K, Laiho R, Tuittila E-S, Vasander H. 2000. Suokasvit - turpeen tekijät           |
| 532 | (Mire vegetation and peat formation). Publications from the Department of Forest Ecology, University     |
| 533 | of Helsinki <b>24:</b> 1-55.   |
| 534 | 13. Bridgham SD, Cadillo-Quiroz H, Keller JK, Zhuang Q. 2013. Methane emissions from                     |

wetlands: biogeochemical, microbial, and modeling perspectives from local to global scales. Global
Change Biol 19:1325-1346.

Microbiology

14. Tveit AT, Urich T, Frenzel P, Svenning MM. 2015. Metabolic and trophic interactions modulate
methane production by Arctic peat microbiota in response to warming. Proc Natl Acad Sci USA 112:
E2507-E2516.

15. Kotsyurbenko OR. 2005. Trophic interactions in the methanogenic microbial community of lowtemperature terrestrial ecosystems. FEMS Microbiol Ecol 53:3-13.

542 16. Glissmann K, Conrad R. 2002. Saccharolytic activity and its role as a limiting step in methane

formation during the anaerobic degradation of rice straw in rice paddy soil. Biol Fertility Soils 35:62-

544 67.

545 17. Conrad R. 1999. Contribution of hydrogen to methane production and control of hydrogen

concentrations in methanogenic soils and sediments. FEMS Microbiol Ecol **28**:193-202.

547 18. Yavitt JB, Seidman-Zager M. 2006. Methanogenic conditions in northern peat soils.

548 Geomicrobiol J **23:**119-127.

549 19. Keller JK, Bridgham SD. 2007. Pathways of anaerobic carbon cycling across an ombrotrophic-

550 minerotrophic peatland gradient. Limnol Oceanogr **52**:96-101.

551 20. Hines ME, Duddleston KN, Rooney-Varga J, Fields D, Chanton JP. 2008. Uncoupling of

acetate degradation from methane formation in Alaskan wetlands: Connections to vegetation

distribution. Global Biogeochem Cycles **22:**GB2017. doi:10.1029/2006GB002903.

554 21. Galand P, Yrjälä K, Conrad R. 2010. Stable carbon isotope fractionation during methanogenesis

in three boreal peatland ecosystems. Biogeosciences **7:**3893-3900.

Microbiology

557 anaerobic carbon mineralization, the efficiency of methane production, and methanogenic pathways in

558 peatlands across an ombrotrophic-minerotrophic gradient. Soil Biol Biochem 54:36-47.

559 23. Pester M, Knorr K, Friedrich MW, Wagner M, Loy A. 2012. Sulfate-reducing microorganisms

560 in wetlands - fameless actors in carbon cycling and climate change. Front Microbiol 3:72.

561 doi:10.3389/fmicb.2012.00072.

562 24. Hausmann B, Knorr K, Schreck K, Tringe SG, del Rio TG, Loy A, Pester M. 2016. Consortia 563 of low-abundance bacteria drive sulfate reduction-dependent degradation of fermentation products in peat soil microcosms. ISME J 10:2365-2375. 564

- 565 25. Corbett JE, Tfaily MM, Burdige DJ, Glaser PH, Chanton JP. 2015. The relative importance of 566 methanogenesis in the decomposition of organic matter in northern peatlands. J Geophys Res: 567 Biogeosci 120:280-293.
- 568 26. Hamberger A, Horn MA, Dumont MG, Murrell JC, Drake HL. 2008. Anaerobic consumers of 569 monosaccharides in a moderately acidic fen. Appl Environ Microbiol 74:3112-3120.
- 570 27. Schmidt O, Hink L, Horn MA, Drake HL. 2016. Peat: home to novel syntrophic species that feed 571 acetate- and hydrogen-scavenging methanogens. ISME J 10:1954-1966.
- 572 28. Schmidt O, Horn MA, Kolb S, Drake HL. 2015. Temperature impacts differentially on the
- 573 methanogenic food web of cellulose-supplemented peatland soil. Environ Microbiol 17:720-734.
- 574 29. Drake HL, Horn MA, Wüst PK. 2009. Intermediary ecosystem metabolism as a main driver of
- 575 methanogenesis in acidic wetland soil. Env Microbiol Rep 1:307-318.

Applied and Environmental Microbioloay 576

| 577 | moderately acidic fen soil. Environ Microbiol 11:1395-1409.  |
|-----|--|
| 578 | 31. Lin X, Tfaily MM, Green SJ, Steinweg JM, Chanton P, Imvittaya A, Chanton JP, Cooper W,         |
| 579 | Schadt C, Kostka JE. 2014. Microbial metabolic potential for carbon degradation and nutrient       |
| 580 | (nitrogen and phosphorus) acquisition in an ombrotrophic peatland. Appl Environ Microbiol 80:3531- |
| 581 | 3540.  |
| 582 | 32. Tveit A, Schwacke R, Svenning MM, Urich T. 2013. Organic carbon transformations in high-       |
| 583 | Arctic peat soils: key functions and microorganisms. ISME J 7:299-311.                             |
| 584 | 33. Hunger S, Gossner AS, Drake HL. 2015. Anaerobic trophic interactions of contrasting methane-   |
| 585 | emitting mire soils: processes versus taxa. FEMS Microbiol Ecol 91:fiv045. doi:                    |
| 586 | 10.1093/femsec/fiv045.   |
| 587 | 34. Juottonen H, Galand PE, Tuittila E-S, Laine J, Fritze H, Yrjälä K. 2005. Methanogen            |
| 588 | communities and Bacteria along an ecohydrological gradient in a northern raised bog complex.       |

30. Wüst PK, Horn MA, Drake HL. 2009. Trophic links between fermenters and methanogens in a

589 Environ Microbiol 7:1547-1557.

590 35. Laine J, Komulainen V, Laiho R, Minkkinen K, Rasinmäki A, Sallantaus T, Sarkkola S,

591 Silvan N, Tolonen K, Tuittila E-S, Vasander H, Päivänen J. 2004. Lakkasuo-guide to a mire

592 ecosystem. Publications from the Department of Forest Ecology, University of Helsinki **31:**1-123.

593 36. Jaatinen K, Tuittila E-S, Laine J, Yrjälä K, Fritze H. 2005. Methane-oxidizing bacteria in a

594 Finnish raised mire complex: effects of site fertility and drainage. Microb Ecol **50**:429-439.

595 37. Korkama-Rajala T, Müller MM, Pennanen T. 2008. Decomposition and fungi of needle litter
596 from slow- and fast-growing Norway spruce (*Picea abies*) clones. Microb Ecol 56:76-89.

Microbiology

597 38. Nicolaisen MH, Bælum J, Jacobsen CS, Sørensen J. 2008. Transcription dynamics of the 598

- 599 in agricultural soil. Environ Microbiol 10:571-579.
- 600 39. Angel R, Claus P, Conrad R. 2012. Methanogenic archaea are globally ubiquitous in aerated soils
- 601 and become active under wet anoxic conditions. ISME J 6:847-862.

602 40. Jaakola L, Pirttilä AM, Halonen M, Hohtola A. 2001. Isolation of high quality RNA from

- 603 bilberry (Vaccinium myrtillus L.) fruit. Mol Biotechnol 19:201-203.
- 604 41. Whiteley AS, Thomson B, Lueders T, Manefield M. 2007. RNA stable-isotope probing. Nat 605 Protoc 2:838-844.
- 606 42. Herlemann DP, Labrenz M, Jürgens K, Bertilsson S, Waniek JJ, Andersson AF. 2011.
- 607 Transitions in bacterial communities along the 2000 km salinity gradient of the Baltic Sea. ISME J 608 **5:**1571-1579.
- 609 43. Sinclair L, Osman OA, Bertilsson S, Eiler A. 2015. Microbial community composition and diversity via 16S rRNA gene amplicons: evaluating the Illumina platform. PloS one 10:e0116955. 610
- 611 doi:10.1371/journal.pone.0116955.
- 612 44. Edgar RC. 2013. UPARSE: highly accurate OTU sequences from microbial amplicon reads. Nat 613 Methods 10:996-998.
- 614 45. Lanzén A, Jørgensen SL, Huson DH, Gorfer M, Grindhaug SH, Jonassen I, Øvreås L, Urich 615 T. 2012. CREST-classification resources for environmental sequence tags. PLoS One 7:e49334.
  - 616 doi:10.1371/journal.pone.0049334.

ied and Environmental Microbiology

Microbiology

617

618

619

620

Nucleic Acids Res 41:D590-6.

# 621 Oakley BB, Parks DH, Robinson CJ, Sahl JW, Stres B, Thallinger GG, Van Horn DJ, Weber CF. 622 2009. Introducing mothur: open-source, platform-independent, community-supported software for 623 describing and comparing microbial communities. Appl Environ Microbiol 75:7537-7541. 624 48. Pinheiro J, Bates D, DebRoy S, Sarkar D, 3R Development Core Team. 2014. nlme: Linear and 625 nonlinear mixed effects models. R package version 3.1-111. http://CRAN.R-project.org/package=nlme. 626 49. R Development Core Team. 2014. R: A language and environment for statistical computing. R 627 Foundation for Statistical Computing, Vienna. http://www.R-project.org. 628 50. Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara R, Simpson GL, 629 Solymos P, Stevens M, Wagner H. 2013. vegan: Community Ecology Package. R package version 630 2.0-10. http://CRAN.R-project.org/package=vegan. 631 51. Robinson MD, McCarthy DJ, Smyth GK. 2010. edgeR: a Bioconductor package for differential 632 expression analysis of digital gene expression data. Bioinformatics 26:139-140. 633 52. McMurdie PJ, Holmes S. 2014. Waste not, want not: why rarefying microbiome data is 634 inadmissible. PLoS Comput Biol 10:e1003531. doi:10.1371/journal.pcbi.1003531. 635 53. Liu J, Zheng C, Song C, Guo S, Liu X, Wang G. 2014. Conversion from natural wetlands to 636 paddy field alters the composition of soil bacterial communities in Sanjiang Plain, Northeast China. 637 Ann Microbiol 64:1395-1403.

46. Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glockner FO. 2013.

The SILVA ribosomal RNA gene database project: improved data processing and web-based tools.

47. Schloss PD, Westcott SL, Ryabin T, Hall JR, Hartmann M, Hollister EB, Lesniewski RA,

Microbiology

638 54. Uz I, Ogram AV. 2006. Cellulolytic and fermentative guilds in eutrophic soils of the Florida 639 Everglades. FEMS Microbiol Ecol 57:396-408. 640 55. Kwon S, Kim B, Kim W, Yoo K, Yoo S, Son J, Weon H. 2008. Paludibacterium yongneupense 641 gen. nov., sp. nov., isolated from a wetland, Yongneup, in Korea. Int J Syst Evol Microbiol 58:190-

642 194.

643 56. Weon H, Kim B, Yoo S, Kwon S, Go S, Stackebrandt E. 2008. Uliginosibacterium gangwonense 644 gen. nov., sp. nov., isolated from a wetland, Yongneup, in Korea. Int J Syst Evol Microbiol 58:131-645 135.

646 57. Sizova MV, Panikov NS, Spiridonova EM, Slobodova NV, Tourova TP. 2007. Novel

647 facultative anaerobic acidotolerant Telmatospirillum siberiense gen. nov. sp. nov. isolated from

648 mesotrophic fen. Syst Appl Microbiol 30:213-220.

649 58. Kodama Y, Watanabe K. 2011. Rhizomicrobium electricum sp. nov., a facultatively anaerobic,

650 fermentative, prosthecate bacterium isolated from a cellulose-fed microbial fuel cell. Int J Syst Evol 651 Microbiol 61:1781-1785.

652 59. Ueki A, Kodama Y, Kaku N, Shiromura T, Satoh A, Watanabe K, Ueki K. 2010.

653 Rhizomicrobium palustre gen. nov., sp. nov., a facultatively anaerobic, fermentative stalked bacterium

654 in the class Alphaproteobacteria isolated from rice plant roots. J Gen Appl Microbiol 56:193-203.

- 655 60. Pankratov TA, Kirsanova LA, Kaparullina EN, Kevbrin VV, Dedysh SN. 2012. Telmatobacter
- 656 bradus gen. nov., sp. nov., a cellulolytic facultative anaerobe from subdivision 1 of the Acidobacteria,
- 657 and emended description of Acidobacterium capsulatum Kishimoto et al. 1991. Int J Syst Evol

658 Microbiol 62:430-437. 31

Microbioloav

659 61. Sizova MV, Panikov NS, Tourova TP, Flanagan PW. 2003. Isolation and characterization of
660 oligotrophic acido-tolerant methanogenic consortia from a *Sphagnum* peat bog. FEMS Microbiol Ecol
661 45:301-315.

662 62. Pester M, Bittner N, Deevong P, Wagner M, Loy A. 2010. A 'rare biosphere'microorganism
663 contributes to sulfate reduction in a peatland. ISME J 4:1591-1602.

664 63. Zinger L, Shahnavaz B, Baptist F, Geremia RA, Choler P. 2009. Microbial diversity in alpine
665 tundra soils correlates with snow cover dynamics. ISME J 3:850-859.

666 64. Sangwan P, Kovac S, Davis KE, Sait M, Janssen PH. 2005. Detection and cultivation of soil
667 verrucomicrobia. Appl Environ Microbiol 71:8402-8410.

668 65. Anders H, Power JF, MacKenzie AD, Lagutin K, Vyssotski M, Hanssen E, Moreau JW, Stott

669 MB. 2015. Limisphaera ngatamarikiensis gen. nov., sp. nov., a thermophilic, pink-pigmented coccus

isolated from subaqueous mud of a geothermal hotspring. Int J Syst Evol Microbiol 65:1114-1121.

671 66. Kant R, van Passel MW, Sangwan P, Palva A, Lucas S, Copeland A, Lapidus A, Glavina del

672 Rio T, Dalin E, Tice H, Bruce D, Goodwin L, Pitluck S, Chertkov O, Larimer FW, Land ML,

673 Hauser L, Brettin TS, Detter JC, Han S, de Vos WM, Janssen PH, Smidt H. 2011. Genome

674 sequence of "Pedosphaera parvula" Ellin514, an aerobic verrucomicrobial isolate from pasture soil. J

675 Bacteriol **193:**2900-2901.

676 67. Chin K, Liesack W, Janssen PH. 2001. Opitutus terrae gen. nov., sp. nov., to accommodate novel

677 strains of the division 'Verrucomicrobia' isolated from rice paddy soil. Int J Syst Evol Microbiol

678 **51:**1965-1968.

Microbiology

679

680

681 a rice paddy field. Int J Syst Evol Microbiol 64:1718-1723. 682 69. Hedlund BP. 2010. Phylum XXIII. Verrucomicrobia phyl. nov., p 795-841. In Krieg NR, Staley 683 JT, Brown DR, Hedlund BP, Paster BJ, Ward NL, Ludwig W, Whitman WB (ed), Bergey's Manual of 684 Systematic Bacteriology, 2nd ed, vol 4. Springer, New York, NY. 685 70. Dedysh SN, Pankratov TA, Belova SE, Kulichevskaya IS, Liesack W. 2006. Phylogenetic 686 analysis and in situ identification of bacteria community composition in an acidic Sphagnum peat bog. Appl Environ Microbiol 72:2110-2117. 687 688 71. Serkebaeva YM, Kim Y, Liesack W, Dedysh SN. 2013. Pyrosequencing-based assessment of the bacteria diversity in surface and subsurface peat layers of a northern wetland, with focus on poorly 689 studied phyla and candidate divisions. PLoS One 8:e63994. doi:10.1371/journal.pone.0063994. 690 691 72. Bergmann GT, Bates ST, Eilers KG, Lauber CL, Caporaso JG, Walters WA, Knight R, 692 Fierer N. 2011. The under-recognized dominance of Verrucomicrobia in soil bacterial communities. 693 Soil Biol Biochem 43:1450-1455. 694 73. Hiras J, Wu Y, Eichorst SA, Simmons BA, Singer SW. 2015. Refining the phylum Chlorobi by 695 resolving the phylogeny and metabolic potential of the representative of a deeply branching, 696 uncultivated lineage. ISME J 10:833-845. 697 74. Dedvsh SN. 2011. Cultivating uncultured bacteria from northern wetlands: knowledge gained and remaining gaps. Front Microbiol 2:184. doi:10.3389/fmicb.2011.00184. 698 699 75. Lin X, Tfaily MM, Steinweg JM, Chanton P, Esson K, Yang ZK, Chanton JP, Cooper W, 700 Schadt CW, Kostka JE. 2014. Microbial community stratification linked to utilization of

68. Qiu Y, Kuang X, Shi X, Yuan X, Guo R. 2014. Terrimicrobium sacchariphilum gen. nov., sp.

nov., an anaerobic bacterium of the class 'Spartobacteria'in the phylum Verrucomicrobia, isolated from

Microbiology

702 Minnesota, USA. Appl Environ Microbiol **80:**3518-3530.

703 76. Lin X, Green S, Tfaily MM, Prakash O, Konstantinidis KT, Corbett JE, Chanton JP, Cooper

704 WT, Kostka JE. 2012. Microbial community structure and activity linked to contrasting

705 biogeochemical gradients in bog and fen environments of the Glacial Lake Agassiz Peatland. Appl

706 Environ Microbiol **78:**7023-7031.

707 77. Basiliko N, Henry K, Gupta V, Moore T, Driscoll B, Dunfield P. 2013. Controls on bacterial and

archaeal community structure and greenhouse gas production in natural, mined, and restored Canadian

709 peatlands. Front Microbiol **4:**215. doi:10.3389/fmicb.2013.00215.

- 710 78. Pankratov TA, Ivanova AO, Dedysh SN, Liesack W. 2011. Bacterial populations and
- environmental factors controlling cellulose degradation in an acidic *Sphagnum* peat. Environ Microbiol **13:**1800-1814.

713 79. Meehan CJ, Beiko RG. 2014. A phylogenomic view of ecological specialization in the

Lachnospiraceae, a family of digestive tract-associated bacteria. Genome Biol Evol 6:703-713.

715 80. Kittelmann S, Pinares-Patiño CS, Seedorf H, Kirk MR, Ganesh S, McEwan JC, Janssen PH.

716 2014. Two different bacterial community types are linked with the low-methane emission trait in sheep.

717 PLoS One 9:e103171. doi:10.1371/journal.pone.0103171.

718 81. Wegner C, Liesack W. 2016. Microbial community dynamics during the early stages of plant

719 polymer breakdown in paddy soil. Environ Microbiol 18:2825–2842.

82. Biddle A, Stewart L, Blanchard J, Leschine S. 2013. Untangling the genetic basis of fibrolytic
specialization by Lachnospiraceae and Ruminococcaceae in diverse gut communities. Diversity 5:627640. doi:10.3390/d5030627.

lied and Environmental Microbioloav

Microbiology

# 723 83. Morales SE, Mouser PJ, Ward N, Hudman SP, Gotelli NJ, Ross DS, Lewis TA. 2006. 724 Comparison of bacterial communities in New England Sphagnum bogs using terminal restriction 725 fragment length polymorphism (T-RFLP). Microb Ecol 52:34-44. 726 84. Ye R, Keller JK, Jin Q, Bohannan BJ, Bridgham SD. 2016. Peatland types influence the 727 inhibitory effects of a humic substance analog on methane production. Geoderma **265**:131-140. 728 85. Cervantes FJ, de Bok FA, Duong-Dac T, Stams AJ, Lettinga G, Field JA. 2002. Reduction of 729 humic substances by halorespiring, sulphate-reducing and methanogenic microorganisms. Environ 730 Microbiol 4:51-57. 731 86. Benz M, Schink B, Brune A. 1998. Humic acid reduction by Propionibacterium freudenreichii and other fermenting bacteria. Appl Environ Microbiol 64:4507-4512. 732 733 87. Lovley DR, Coates JD, Blunt-Harris EL, Phillips EJ, Woodward JC. 1996. Humic substances 734 as electron acceptors for microbial respiration. Nature 382:445-448. 735 88. Lovley D. 2013. Dissimilatory Fe (III)-and Mn (IV)-reducing prokaryotes, p 287-308. In 736 Rosenberg E, DeLong EE, Lory S, Stackebrandt E, Thompson F (ed), The prokaryotes - prokaryotic physiology and biochemistry, 4th ed. Springer, New York, NY. 737 738 89. Thrash JC, Coates JD. 2010. Phylum XVII. Acidobacteria phyl. nov., p. 725-735. In Krieg NR, 739 Staley JT, Brown DR, Hedlund BP, Paster BJ, Ward NL, Ludwig W, Whitman WB (ed), Bergey's 740 Manual of Systematic Bacteriology, 2nd ed, vol 4. Springer, New York, NY. 741 90. Coupland K, Johnson DB. 2008. Evidence that the potential for dissimilatory ferric iron reduction

is widespread among acidophilic heterotrophic bacteria. FEMS Microbiol Lett **279:**30-35.

Microbioloav

743

# cycling bacteria in acidic sediments. Appl Environ Microbiol 76:8174-8183. 92. Lipson DA, Haggerty JM, Srinivas A, Raab TK, Sathe S, Dinsdale EA. 2013. Metagenomic insights into anaerobic metabolism along an Arctic peat soil profile. PLoS One 8:e64659. doi:10.1371/journal.pone.0064659. 93. Reiche M, Torburg G, Kusel K. 2008. Competition of Fe(III) reduction and methanogenesis in an acidic fen. FEMS Microbiol Ecol 65:88-101. 94. Kulichevskaya IS, Suzina NE, Rijpstra WIC, Damsté JSS, Dedysh SN. 2014. Paludibaculum

fermentans gen. nov., sp. nov., a facultative anaerobe capable of dissimilatory iron reduction from

91. Lu S, Gischkat S, Reiche M, Akob DM, Hallberg KB, Kusel K. 2010. Ecophysiology of Fe-

subdivision 3 of the Acidobacteria. Int J Syst Evol Microbiol **64:**2857-2864.

95. Gupta V, Smemo KA, Yavitt JB, Fowle D, Branfireun B, Basiliko N. 2013. Stable isotopes
reveal widespread anaerobic methane oxidation across latitude and peatland type. Environ Sci Technol

**47:**8273-8279.

# 756 96. Horn MA, Matthies C, Kusel K, Schramm A, Drake HL. 2003. Hydrogenotrophic

methanogenesis by moderately acid-tolerant methanogens of a methane-emitting acidic peat. Appl
Environ Microbiol 69:74-83.

# 759 97. Wang X, Sharp CE, Jones GM, Grasby SE, Brady AL, Dunfield PF. 2015. Stable-isotope

probing identifies uncultured planctomycetes as primary degraders of a complex heteropolysaccharide

761 in soil. Appl Environ Microbiol **81:**4607-4615.

762 98. Schellenberger S, Kolb S, Drake HL. 2010. Metabolic responses of novel cellulolytic and

saccharolytic agricultural soil bacteria to oxygen. Environ Microbiol 12:845-861.

Microbiology

# 764 99. Podosokorskaya OA, Bonch-Osmolovskaya EA, Novikov AA, Kolganova TV, Kublanov IV.

765 2013. Ornatilinea apprima gen. nov., sp. nov., a cellulolytic representative of the class Anaerolineae.

766 Int J Syst Evol Microbiol 63:86-92.

- 100. Wrighton KC, Castelle CJ, Wilkins MJ, Hug LA, Sharon I, Thomas BC, Handley KM,
- 768 Mullin SW, Nicora CD, Singh A. 2014. Metabolic interdependencies between phylogenetically novel
- fermenters and respiratory organisms in an unconfined aquifer. ISME J 8:1452-1463.

# 101. Kantor RS, Wrighton KC, Handley KM, Sharon I, Hug LA, Castelle CJ, Thomas BC,

- 771 Banfield JF. 2013. Small genomes and sparse metabolisms of sediment-associated bacteria from four
- candidate phyla. MBio **4:**e00708-13. doi:10.1128/mBio.00708-13.
- 102. Nelson WC, Stegen JC. 2015. The reduced genomes of Parcubacteria (OD1) contain signatures
- of a symbiotic lifestyle. Front Microbiol **6:**713. doi:10.3389/fmicb.2015.00713.

categories of >2.5%, 0.25-2.5%, and 0.025-0.25% of reads) identified as enriched in the  $^{13}$ C fractions

777 of <sup>13</sup>C-cellobiose labelled peat from oligotrophic and mesotrophic fen.

| Classification of OTUs  |                    | 0<br>>2.5%     | ligotrop<br>>0.25% | hic <sup>a</sup><br>>0.025% | mesot<br>>2.5% | trophic <sup>a</sup><br>>0.25% | >0.025% | shared |
|-------------------------|--------------------|----------------|--------------------|-----------------------------|----------------|--------------------------------|---------|--------|
| Firmicutes              | Clostridiaceae     | 3              | 2 <sup>1</sup>     | 3 <sup>1</sup>              | 8              | 7 <sup>1</sup>                 | 2       |        |
| Firmicules              | Lachnospiraceae    | 5              | Z                  | 3                           | 0              | /                              | Z       |        |
|                         | Ruminococcaceae    |                |                    |                             |                | 3                              | 1       | 13     |
|                         | Veillonellaceae    |                | $7^{3}$            | $2^{2}$                     | 1              | $9^{2}$                        | 7       | 15     |
|                         | unknown family     |                | /                  | 2                           | 1              | 2                              | 2       |        |
| Betaproteobacteria      | Burkholderiales    |                | 1                  |                             |                | 2                              | 2       |        |
| Бешргогеобастегна       | Neisseriales       | 3 <sup>1</sup> | 1                  | 4                           | 2              |                                | 2       |        |
|                         | Rhodocyclales      | 3              | 2                  | 2                           | $\frac{2}{2}$  | 1                              | 3       | 5      |
|                         | Myxococcales       | 5              | 2                  | 1                           | 2              | 1                              | 5       |        |
| Alphaproteobacteria     | Rhizobiales        | 1              | 3 <sup>1</sup>     | 3                           |                |                                | 1       |        |
| I I                     | Rhodospirillales   |                | 4                  | 3                           | 1              | 3                              | 3       | 5      |
| Deltaproteobacteria     | Anaeromyxobacter   |                |                    | 1                           |                |                                |         | 0      |
| Gammaproteobacteria     | Tolumonas          |                | 1                  |                             | 2              |                                |         | 1      |
| Acidobacteria           | Acidobacteriaceae  | 1              | 13                 | 4                           |                |                                |         |        |
|                         | Holophagaceae      |                | 3 <sup>1</sup>     | 3                           |                | 1                              | $1^{1}$ | 2      |
|                         | group 3            |                | 2                  |                             |                |                                |         |        |
| Verrucomicrobia         | subphylum 3        |                | 3                  | 11                          |                |                                |         |        |
|                         | Spartobacteria     |                | 1                  | 1                           |                |                                |         | 0      |
|                         | Opitutae           |                | 2                  | 2                           |                |                                |         |        |
| Spirochaetes            | Spirochaetales     |                | 5                  | 5                           |                |                                | 1       | 0      |
| Bacteroidetes           | Bacteroidales      |                |                    | 2                           |                | $1^{1}$                        | $1^{1}$ |        |
|                         | Sphingobacteriales |                |                    |                             |                |                                | 1       | 0      |
|                         | other/unknown      |                | 2                  | 3                           |                |                                |         |        |
| Candidate division BRC1 |                    |                | 1                  |                             |                |                                |         | 0      |
| Actinobacteria          | Coriobacteriia     |                | $1^{1}$            |                             |                |                                |         | 0      |
| Armatimonadetes         | group 1            |                |                    | 2                           |                |                                |         | 0      |
| Planctomycetes          | Phycisphaerae      |                |                    | 1                           |                |                                |         | 0      |
| Chlorobi                | Ignavibacteria     |                |                    | 1                           |                |                                |         | 0      |
| Melainabacteria         |                    |                |                    |                             |                |                                | 1       | 0      |

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<sup>a</sup>The superscripts indicate the number of OTUs that became more <sup>13</sup>C-enriched from day 7 to day 14.

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## 781 Figure legends

Fig. 1. A) <sup>13</sup>C-enrichment of CH<sub>4</sub> and CO<sub>2</sub> and B) <sup>13</sup>C-CH<sub>4</sub> and <sup>13</sup>C-CO<sub>2</sub> production rates during labeling experiment with [<sup>13</sup>C]cellobiose in oligotrophic and mesotrophic peat at three time points (n=3, mean $\pm$  SD).

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**Fig. 2.** Non-metric multidimensional scaling (NMDS) plot of bacterial community based on 16S rRNA gene sequencing in original peat and in the heavy ( $^{13}$ C) and light ( $^{12}$ C) stable isotope fractions of oligotrophic and mesotrophic peat. Stress = 0.084.

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790 Fig. 3. Log<sub>2</sub> fold changes of read abundances showing the taxonomic distribution of bacterial

791 operational taxonomic units (OTUs, 150 most abundant) in heavy (<sup>13</sup>C) and light (<sup>12</sup>C) fractions at A)

day 7 and B) day 14. Reads more abundant in mesotrophic peat, x axis < 0; in oligotrophic peat, x axis

793 > 0; in <sup>13</sup>C fractions, y axis > 0; in <sup>12</sup>C fractions, y axis < 0. Solid grey line separates OTUs at least 2-

fold more abundant than in the other peat type or fraction. Fold changes were calculated from means of

three replicate incubations using data subsampled to 13226 reads for each sample. Symbol size is based

on the read number of an OTU over all the samples at each time point.





