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Effects of Initial Age Structure of Managed Norway Spruce Forest Area on Net Climate Impact of Using Forest Biomass for Energy

Kilpeläinen A

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1 **Effects of initial age structure of managed Norway spruce forest area on net climate impact of**
2 **using forest biomass for energy**

3
4 ¹Kilpeläinen, A., ¹Strandman, H., ²Grönholm, T., ¹Ikonen, V.-P., ¹Torssonen, P., ¹Kellomäki, S.,
5 and ¹Peltola, H.

6
7 ¹University of Eastern Finland, Faculty of Science and Forestry, School of Forest Sciences
8 P.O. Box 111, FI-80101 Joensuu, Finland

9
10 ²University of Helsinki, Faculty of Science, Department of Physics, Division of Atmospheric Sciences
11 P.O. Box 48, FI-00014 Helsinki, Finland

12
13 Corresponding author: e-mail: antti.kilpelainen@uef.fi; Tel. +358 50 382 3263

14
15
16 **Abstract**

17
18 We investigated how the initial age structure of a managed, middle boreal (62° N), Norway spruce
19 dominated (*Picea abies* L. Karst.) forest area affects the net climate impact of using forest biomass
20 for energy. The model-based analysis used a gap-type forest ecosystem model linked to a life cycle
21 assessment (LCA) tool. The net climate impact of energy biomass refers to the difference in annual
22 net CO₂ exchange between the biosystem using forest biomass (logging residues from final felling)
23 and the fossil (reference) system using coal. In the simulations over the 80-year period, the alternative
24 initial age structures of the forest areas were: (i) skewed to the right (dominated by young stands), (ii)
25 normally distributed (dominated by middle-aged stands), (iii) skewed to the left (dominated by mature
26 stands), and (iv) evenly distributed (same share of different age classes). The effects of management
27 on net climate impacts were studied using current recommendations as a baseline with a fixed rotation
28 period of 80 years. In alternative management scenarios, the volume of the growing stock was
29 maintained 20% higher over the rotation compared to the baseline, and/or nitrogen fertilization was
30 used to enhance carbon sequestration. According to the results, the initial age structure of the forest
31 area affected largely the net climate impact of using energy biomass over time. An initially right-
32 skewed age structure produced the highest climate benefits over the 80-year simulation period, in
33 contrast to the left-skewed age structure. Furthermore, management that enhanced carbon
34 sequestration increased the potential of energy biomass to replace coal; reducing CO₂ emissions and
35 enhancing climate change mitigation.

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37
38 **Keywords:** bioenergy, climate impact, forest biomass, forest management, radiative forcing,
39 substitution

40
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51 **Introduction**

52

53 An understanding of the dynamics of carbon flows in managed forests is needed to increase their
54 capacity to remove and store carbon dioxide (CO₂) outside the atmosphere [1]. Management and
55 harvest cycles over a rotation affect the carbon sequestration potential of the growing stock, as well
56 as the potential for biomass harvest. The potential of biomass for replacing fossil-based materials
57 (e.g. concrete and plastics) and fossil energy (e.g. coal and oil) is further affected by the harvest
58 intensity of the biomass (e.g. saw log, pulp wood, energy biomass). All of these factors affect the
59 possibilities for mitigating emissions through increasing carbon sequestration in ecosystems, and
60 using forest biomass in place of fossil fuels. In this context, the timing of emissions has an impact on
61 the climate benefits of biomass production and utilization when compared to the use of fossil
62 counterparts, and consequently on efficient management measures for climate change mitigation [e.g.
63 1–6].

64 A forested area initially dominated by mature stands provides the maximum potential for
65 timber, whereas areas with a normal age distribution of stands, or those dominated by young stands
66 provide the highest carbon sequestration in the short term [7, 8]. Harvested biomass produces the
67 highest climate benefits when it is used in replacing both fossil fuels and/or fossil-based materials,
68 while maintaining sustainable carbon sequestration of forests [9–11]. In the case of energy biomass
69 (e.g. logging residues), the initial level of carbon sequestration in a forested area critically affects the
70 mitigation potential. This is because the amount of CO₂ released into the atmosphere during the
71 combustion of energy biomass may be higher per energy unit than that emitted when using fossil
72 fuels (e.g. coal), temporarily increasing the amount of carbon in the atmosphere [12, 13]. However,
73 the temporary increase in CO₂ is offset by carbon sequestration after regrowth of the stand, in a time
74 frame that is dependent on forest management practices and harvest intensity [11]. Therefore, the

75 joint impacts of CO₂ release and carbon sequestration on total emission mitigation are highly sensitive
76 to the initial conditions of the studied biosystem [14].

77 In general, the carbon uptake of a forest stand peaks earlier than the point when the maximum
78 carbon stock (carbon in trees and soil) is achieved. Therefore, forest area initially dominated by young
79 stands have been found to produce the lowest emissions for energy biomass utilization when rotation
80 lengths of 60–80 years are used [7]. In turn, rotation lengths of 120 years in a forest area with initially
81 normally distributed age structure have been found to produce the lowest emissions [7]. Intensive
82 management of a single stand, including the use of nitrogen fertilizers, improved planting material,
83 and high stocking density, also increases the climate benefits of using energy biomass compared to
84 the use of fossil-based energy under alternative rotation periods, due to increased carbon sequestration
85 [9, 15–20]. However, the potential of forest management to increase carbon sequestration within a
86 certain time frame also depends on the age structure of a forest area. This further affects the temporal
87 distribution of the future climate benefits and CO₂ emissions of energy biomass utilization.

88 In this context, the effects of the alternative initial age structures and management of a Norway
89 spruce (*Picea abies* L. Karst.) forest area in the middle boreal zone (62° N) on the net climate impact,
90 and its timing when using forest biomass for energy, were quantified. A model-based analysis using
91 a gap-type forest ecosystem model, linked to a life cycle assessment (LCA) tool was used. The net
92 climate impact of energy biomass is defined as the annual difference in net CO₂ exchange between
93 the biosystem using forest biomass (logging residues from final felling), and the fossil (reference)
94 system using coal, over an 80-year period. The net climate impact of energy biomass was expressed
95 in terms of carbon neutrality and relative radiative forcing. Four initial age structures of a forest area
96 were used in initializing the simulations: (i) skewed to the right, (ii) normally distributed, (iii) skewed
97 to the left, and (iv) evenly distributed. To determine the effects of management, current forest
98 management recommendations (baseline management, without harvesting of logging residues) were
99 used as a reference. In alternative management scenarios, stocking density (basal area of the stands)

100 was maintained at a level 20% higher over the rotation than in the baseline, and/or nitrogen
101 fertilization was used to enhance carbon sequestration and energy biomass production.

102

103 **Materials and methods**

104

105 **Outlines for calculating net climate impacts between biosystems and fossil system**

106 The net climate impact for energy biomass production and utilization was calculated by comparing
107 the use of logging residues from final fellings (including the top part of the stem, branches, 70% of
108 needles, stumps, and coarse roots) to the use of fossil coal. The calculation was carried out for model
109 forest areas consisting of 80 single Norway spruce stands (each covering 1 hectare) and by using
110 alternative initial age structures for these areas at the beginning of the 80-year simulation period (**Fig.**
111 **1**). The alternative initial stand age structures were used to quantify in which initial conditions, and
112 within which time horizon energy, biomass production and its utilization for fossil coal substitution
113 gave the highest net climate impact. The alternative initial age structures were created by using the
114 even age class distribution, since each stand represented one age class in this case. The tree stand
115 characteristics and amount of initial soil organic matter for the stands of the even age structure were
116 created based on an 80-year simulation using current management recommendations (see also [21]).
117 Current management practices were used to make the initial conditions of the forested areas
118 comparable with the alternative management scenarios.

119 The initial values for the alternative age structures were derived by multiplying the values for
120 the tree stand characteristics and soil organic matter of the even age structure by the relative
121 frequencies of the stands in the forest area shown in **Fig. 1**. The initial age structure of stands in the
122 forest area used in the calculations were (i) right-skewed (most of the stands are young), (ii) normally
123 distributed (most of the stands are middle-aged), (iii) left-skewed (most of the stands are mature), and
124 (iv) even age class distribution (each stand represents one age class and one stand is harvested every

125 year) (**Fig. 1**). The mean ages of the stands for right-skewed, normally distributed, and left-skewed
126 structures were 10, 40, and 70 years, and the standard deviations of the stand ages were 40, 30, and
127 40 years, respectively.

128 In the comparison of the biosystem and fossil system, the (current) baseline management was
129 used as a reference forest management for the fossil system (**Fig. 2**). Baseline management included
130 only timber (sawlogs and pulpwood) harvest—logging residues were not harvested. In the biosystem,
131 alternative management scenarios deviating from the baseline management were used to study the
132 sensitivity of the climate change mitigation potential of energy biomass under the alternative initial
133 age structures. The alternative management scenarios aimed to enhance the carbon sequestration and
134 energy biomass production of the forests in the area. When energy biomass was used, it always
135 replaced fossil coal in the calculations. The annual quantity of energy was assumed to be equal in
136 both systems, so that the use of fossil energy followed the amount of energy in the harvestable energy
137 biomass in the biosystem. When fossil energy (coal) was utilized, the logging residues were left on
138 the site to decay, emitting carbon to the atmosphere. The energy content used for energy biomass was
139 8.1 GJ Mg^{-1} (40% moisture content of wood) [21, 22, 23], and the CO_2 mass emission factor used for
140 the combustion of coal was 93.3 kg GJ^{-1} [24].

141

142 **Calculation of net climate impact for the production and utilization of energy biomass**

143

144 The net climate impact of the production and utilization of energy biomass (I_{NET}) in comparison to
145 the use of coal was indicated by the differences in: (i) annual net CO_2 exchange, (ii) carbon neutrality,
146 and (iii) radiative forcing. In each case, the differences were related to the CO_2 exchange (C_{net})
147 between the alternative initial stand age structures and management scenarios in the biosystem and
148 the baseline (reference) scenario in the fossil system (**Eq. 1**):

149

$$I_{NET} = (a + (a - d) \times c) - (b) \quad (1)$$

151

152 In Equation 1, a is the C_{net} ($\text{g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$) (**Eq. 2**) of the biosystem under baseline management with
 153 the use of energy biomass, b is the C_{net} ($\text{g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$) of the fossil system under baseline (reference)
 154 management with the use of coal, c is the share of energy biomass from the total harvested biomass
 155 (between 0 and 1), and d is the net CO_2 exchange of alternative management scenarios ($\text{g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$).
 156 The enhanced carbon sequestration in the alternative management scenarios was partitioned by the
 157 share of the energy biomass from total harvested biomass over the rotation period. Partition was done
 158 to allow the biosystem to benefit only from the share that belongs to energy biomass.

159 The calculation of annual C_{net} ($\text{g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$) (**Eq. 2**) is based on the carbon uptake during
 160 growth (C_{seq}), carbon emissions from decaying soil organic matter (C_{decomp}), and the combustion of
 161 either energy biomass in the biosystem (C_{harv}) or coal in the fossil system (C_{coal}). The release of carbon
 162 during the combustion of energy biomass was assumed to take place immediately after harvesting.

163

$$C_{net} = C_{seq} + C_{decomp} + C_{harv} \vee C_{coal} \quad (2)$$

165

166 The carbon neutrality (CN) of energy biomass production and utilization was calculated by comparing
 167 the net CO_2 exchange (C_{net}) of using energy biomass to the CO_2 exchange of the utilization of coal
 168 (**Eq. 3**), in $\text{kg CO}_2 \text{ MWh}^{-1}$.

169

$$\text{CN} = \frac{\text{FOSSIL SYSTEM} - \text{BIOSYSTEM}}{\text{FOSSIL SYSTEM}} \quad (3)$$

171

172 Radiative forcing (RF) refers to the change in the balance of incoming and outgoing energy in the
 173 Earth-atmosphere system. A negative RF cools the surface of the Earth, whereas a positive RF warms

174 the surface. The net climate impact (I_{NET}) was converted into the instantaneous RF (Wm^{-2}) using **Eq.**
175 **4–6** [25].

176

$$177 \quad \Delta RF = \alpha \ln\left(\frac{C}{C_0}\right) \quad (4)$$

$$178 \quad \Delta C(t) = \int_0^t E(\tau) f(t - \tau) d\tau \quad (5)$$

$$179 \quad f_{CO_2}(t) = a_0 + \sum_{i=1}^3 a_i e^{-t/\tau_i} \quad (6)$$

180

181 In **Eq. 4**, the parameter α has the value $5.35 Wm^{-2}$, C_0 is the reference concentration of CO_2 in the
182 atmosphere, and C is the prevailing CO_2 concentration, both in ppm. In **Eq. 5**, $\Delta C(t)$ is the change in
183 the CO_2 concentration (ppm), τ is time, t is the time period in question, and E is the change in CO_2
184 concentration in the atmosphere (expressed in kg relative to the total mass of atmosphere in kg). In
185 **Eq. 6**, the function f is the lifetime function of CO_2 , indicating the decay of a pulse of CO_2 with time,
186 where a_0 , a_1 , a_2 , and a_3 are parameters with the values 0.217, 0.259, 0.338, and 0.186. Similarly, the
187 values of the parameters τ_1 , τ_2 , and τ_3 are 172.9, 18.5, and 1.2, respectively [26].

188 For the final analysis, cumulative radiative forcing (CRF) (an integral of instantaneous radiative
189 forcing over time) was calculated over the 80-year simulation period for the areas with alternative
190 initial stand age structures and management scenarios. The same scenarios of the atmospheric CO_2
191 concentrations were used in the forest growth simulations and the radiative forcing calculations.

192

193 **Model simulations needed to calculate an annual net CO_2 exchange for the management**
194 **scenarios**

195

196 *Outline of the forest ecosystem model*

197

198 The outputs of ecosystem model simulations [27] were utilized to estimate the annual net CO₂
199 exchanges (C_{net}) in the systems by using a life cycle assessment (LCA) tool [28]. The forest ecosystem
200 model SIMA [29, 27, 15, 16] was used to simulate the forest growth and development of the tree
201 stands under the influences of temperature; the availability of light, soil water, and nitrogen; and the
202 CO₂ concentration in the atmosphere. Trees may die randomly, but the probability of death increases
203 with a decline in growth. In the simulations, organic matter in litter and dead trees ended up in the
204 soil where it decayed, releasing CO₂ and nitrogen. In decomposition, carbon emissions originated
205 from new litter, old litter, and humus on the site. The old soil organic matter (SOM) was 67 Mg ha⁻¹
206 for the initial medium fertile stand (*Myrtillus* type) [26]. New litter and humus made up the SOM
207 generated during the simulations. The dynamics of available nitrogen were determined by the amount
208 of nitrogen released and immobilized in the decomposition of soil organic matter. Annual nitrogen
209 deposition was set at 10.0 kg ha⁻¹ [30]. The decomposition rate parameters for the litter and humus of
210 Norway spruce are shown in [31]. In simulations, the current climate (temperature, precipitation)
211 during the period 1971–2000 [32, 33] was used. The atmospheric CO₂ concentration was 372 ppm.

212 213 *Alternative management and harvesting scenarios used in the simulations*

214
215 The alternative management and harvesting scenarios used in the simulations are shown in **Table 1**.
216 In the baseline management (BT), only timber was harvested and the current forest management
217 recommendations were followed [34, 35]. The timing and frequency of thinning over a fixed rotation
218 period of 80 years were determined based on the thresholds for basal area at given dominant heights.
219 Energy biomass was harvested only in the final felling of the stands at the end of the 80-year rotation
220 (BNR scenarios). Energy biomass included logging residues (top part of stem, branches, and 70% of
221 needles) with stumps and coarse roots [34]. In alternative management scenarios, 20% higher basal
222 area thresholds were applied for thinning to maintain higher stocking density (basal area of the stand)

223 over a rotation compared to the baseline management (BNR20). In addition, nitrogen fertilization
224 (150 kg ha⁻¹ at fixed years of 50 and 65) was applied (BNRF and BNR20F). After final felling, the
225 site was planted with 2500 Norway spruce seedlings per hectare and managed following the
226 alternative management scenarios.

227

228 **Results**

229

230 *Energy biomass production*

231

232 The production of energy biomass was the highest for the initially left-skewed and the lowest for the
233 initially right-skewed age structure at the beginning of the study period. The initially normally-
234 distributed age structure reached energy biomass production similar to that of the initially evenly-
235 distributed age structure 40 years from the start of the simulation in all the management scenarios.
236 Mean cumulative energy biomass production (expressed in energy units) over the whole simulation
237 period, and in all the initial age structures, was the highest when maintaining 20% higher stocking
238 density than in the baseline, and using nitrogen fertilization in the forest area (2.55 MWh ha⁻¹yr⁻¹)
239 (**Fig. 3**). When maintaining 20% higher stocking density, energy biomass production was 2.50 MWh
240 yr⁻¹ and when using nitrogen fertilization, it was 2.06 MWh ha⁻¹yr⁻¹ over the 80-year simulation
241 period. Under the baseline scenario with energy biomass harvesting (BNR), the energy biomass
242 production was at its lowest (2.02 MWh ha⁻¹yr⁻¹).

243

244 *Net climate impact in terms of difference in annual net CO₂ exchanges (C_{net}) between the biosystem*
245 *and fossil system*

246

247 The net climate impact (I_{NET}) of the energy biomass production and utilization is shown in **Fig. 4** as
248 a cumulative mean difference in net CO_2 exchange (C_{net}) between the biosystem and fossil system.
249 Initially, the highest values of net climate impact were for left-skewed and the lowest ones for right-
250 skewed age structure. The net climate impact was highest (lowest climate benefit) in the baseline
251 management scenario ($-600 \text{ kg } CO_2 \text{ ha}^{-1} \text{ yr}^{-1}$) and lowest (highest climate benefit) in all the initial age
252 structures when maintaining 20% higher stocking density with nitrogen fertilization ($-1200 \text{ kg } CO_2$
253 $\text{ha}^{-1} \text{ yr}^{-1}$) over the simulation period.

254

255 *Net climate impact in terms of carbon neutrality (CN) and avoided CO_2 emissions between the*
256 *biosystem and fossil system*

257

258 The cumulative mean CN values for the energy biomass were 0.87, 1.14, 1.14, and 1.39 over the
259 whole simulation period in baseline (BNR), nitrogen fertilization (BNRF), 20% higher stocking
260 density (BNR20), and nitrogen fertilization and 20% higher stocking density (BNR20F), respectively
261 (**Fig. 5**). CN values were initially highest (highest climate benefits) for the right-skewed age structure
262 and the lowest (lowest climate benefits) for the left-skewed age structure.

263

264 The use of energy biomass instead of coal avoided CO_2 emissions of 257–424 kg MWh^{-1} over the
265 study period (**Table 2**). The highest reductions were found when maintaining 20% higher stocking
266 density, using nitrogen fertilization, and in the baseline management. Maintaining 20% higher
267 stocking density with nitrogen fertilization increased net carbon sequestration and energy biomass
268 production the most (**Fig. 3 and 4**), but could not reach the highest amount of emissions avoided per
269 produced MWh over the whole study period. The lowest avoided emissions per MWh were for the
270 even age structure, followed by the normal, left-skewed, and right-skewed initial age structures
271 (**Table 2**).

272

273 *Net climate impact in terms of radiative forcing (RF) between the biosystem and fossil system*

274

275 The net climate impacts (I_{NET}) of the utilization of energy biomass compared to the use of fossil coal
276 under alternative initial age class structures and management scenarios are shown in **Fig. 6** in terms
277 of radiative forcing. Generally, negative forcing (cooling climate impact) was found for the
278 production and utilization of energy biomass as compared to coal, up to -5.8 nWm^{-2} . The lowest
279 cooling impact was found for the initially left-skewed age structure. By using baseline management
280 (BNR) for this structure, the climate benefits realized at the end of the simulation. In the initially left-
281 skewed age structure, the use of nitrogen fertilization increased the net climate impact, and the
282 radiative forcing was lower after 40 years from the start of the simulation period than when coal was
283 used. With this same age structure, when maintaining 20% higher stocking density, with or without
284 nitrogen fertilization, the net climate impact was lower after the first 10 years than when coal was
285 used. Whereas with the other initial age structures, the climate benefits were gained after a few years
286 from the start of the use of energy biomass.

287

288 **Discussion**

289

290 In this study, the effects of initial stand age structures and management of Norway spruce in a forest
291 area in the middle boreal zone (62° N) on the net climate impact of energy biomass production and
292 utilization was analyzed. Forest ecosystem model [27] simulations linked to a life cycle assessment
293 (LCA) tool [28] were used. The net climate impact was calculated by comparing the CO_2 exchange
294 of the use of logging residues harvested from final fellings to the use of fossil coal, and was further
295 expressed in terms of carbon neutrality, average amount of avoided CO_2 emissions over the study
296 period, and relative radiative forcing. The energy biomass produced in the forest area always replaced

297 fossil coal in the calculations. To determine the effects of management, current forest management
298 recommendations were used as a reference. In the biosystem, alternative management scenarios
299 included 20% higher stocking density over the rotation and/or nitrogen fertilization to increase carbon
300 sequestration and energy biomass production.

301 The net climate impact of energy biomass production and utilization over the simulation period
302 was the lowest (highest cooling impact) if the stand age structure of the forest area was initially right-
303 skewed, and the highest if it was left-skewed. This was due to differences in net carbon sequestration
304 and timing of emissions from energy biomass utilization, in comparison with the fossil system under
305 the alternative age structures. With the initially left-skewed age structure, and under baseline
306 management, the carbon neutrality ($CN = 1$) was not obtained over the study period, but the use of
307 energy biomass produced lower emissions than fossil coal ($CN > 0$). In this case, over the initial 19
308 years of the simulation, the emissions from energy biomass were lower than when using coal, which
309 was in line with studies showing that logging residues from final felling have lower climate impacts
310 than coal about 20 years after use e.g. [36–39, 11]. In this case, maintaining 20% higher stocking
311 density and using nitrogen fertilization, a 60% increase in carbon neutrality of energy biomass could
312 be obtained over the study period compared to the baseline management.

313 Over the whole study period, the largest quantity of emissions avoided (per produced MWh)
314 was also found for the right-skewed initial age structure when maintaining a 20% higher stocking
315 density. Maintaining a 20% higher stocking density and using fertilization increased both carbon
316 sequestration and energy biomass production, but did not result in the highest amount of avoided
317 emissions per energy unit when considered over the whole study period. This was especially the case
318 for the even initial age structure in which the management measures (e.g. fertilization) taking place
319 in the later stages of stand development mostly increased the energy biomass production rather than
320 net carbon sequestration. However, the fastest occurrence of net climate impact was found for the
321 initially right-skewed age structure in which management regimes could also increase carbon

322 sequestration efficiently in the early stages of stand development. For the initially normal and left-
323 skewed age structures, the intensified management regimes produced lower amounts of avoided
324 emissions compared to the baseline management over the whole study period.

325 Based on our results for radiative forcing, the climate benefits of using energy biomass were
326 realized later than indicated by carbon neutrality (CN). As the radiative forcing calculation takes into
327 account the accumulation of carbon released into the atmosphere, any change in carbon concentration
328 in the atmosphere at the beginning of the simulation period is emphasized. The cumulative radiative
329 forcing under baseline management, for example, became negative after 75 years in the initially left-
330 skewed forest area, and after 13 years in the initially right-skewed forest area (Figure 6). Use of the
331 large amount of energy biomass from the initially left-skewed forest area caused a delay in recovery
332 of emissions from using coal, but finally resulted in a cooling climate impact, compared to the fossil
333 system (e.g. [12, 40–42]). However, under other initial age structures, radiative forcing became
334 negative a few years after the start of the simulation. Therefore, the interpretation of the results from
335 fully dynamic approaches and from the static ones (e.g. global warming potential, GWP) should be
336 done with care when assessing net climate impacts of energy biomass from forest ecosystems.

337 Regardless of the initial age structure of the forest area, the use of higher stocking density and/or
338 nitrogen fertilization could enhance the climate benefits of energy biomass. However, forest
339 management also enhanced climate benefits most in the forested area with an initially right-skewed
340 or even age-class structure. Previously, it has generally been observed that forest management could
341 affect largely the climate benefits when using biomass for energy [9, 43]. Forest fertilization has been
342 found to reduce carbon emissions and to increase the availability of primary energy in the integrated
343 production of timber and energy biomass [44]. Decreases in carbon emissions have also been found
344 when emissions for nitrogen fertilizers were included in the analysis [15]. However, the availability
345 of nutrients after the harvesting of logging residues may decrease, affecting the long-term
346 productivity of the forest ecosystem [45, 46]. The whole tree harvesting of energy biomass from

347 thinning has been found to reduce the volume growth of Norway spruce stands (6%) during the first
348 10 years after thinning [47]. However, in our case the logging residues were harvested only from final
349 felling. On the other hand, the nitrogen fertilization may also decrease microbial activity and decrease
350 decomposition of organic matter in the soil [48].

351 In this study, higher rates of carbon sequestration could also shorten the time during which the
352 emissions from the comparator fossil system were recovered, regardless of initial age structure.
353 Previously, the lowest emissions for energy biomass in the integrated production of energy biomass
354 and timber have been found over rotation lengths of 60–80 years [7]. The cooling climate impact
355 (negative radiative forcing) started under the baseline management 35 years earlier when nitrogen
356 fertilization was used in the forest area with a left-skewed age structure. With this same initial age
357 structure, using 20% higher stocking density and nitrogen fertilization, the cooling impact was evident
358 after 10 years. In general, evaluation of the climate benefits of energy biomass based on radiative
359 forcing has been stated to produce overestimates over short-term periods, compared to estimations
360 expressed as global surface temperature change [38]. On the other hand, if the effects of climate
361 change are included in the long-term analyses of the carbon sequestration of the forest ecosystems,
362 the efficiency of energy biomass in replacing fossil coal may decrease due to an increase in CO₂
363 concentration in the reference climate scenario [21].

364 In conclusion, the results of this study show that the initial stand age structure of the forest area
365 plays an important role in determining the potential of using energy biomass in climate change
366 mitigation in boreal conditions over different time spans. The initial age structure affects the
367 possibility of using forest management to increase carbon sequestration and energy biomass
368 production, and consequently the potential of energy biomass in replacing fossil coal. The results of
369 this study also show that forest management that increases carbon sequestration recovers the initially
370 higher unit emissions of the energy biomass faster compared to those of fossil coal. Our analysis
371 considered only the net climate impacts for using energy biomass. Therefore, future work should also

372 consider the impacts of using wood products in fossil material substitution. Further analyses are still
373 needed to define efficient management of alternative biomass-based production systems for climate
374 change mitigation, as well as under a changing climate.

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397 **References**

398

- 399 [1] Canadell JG and Raupach MR (2008) Managing forests for climate change mitigation. *Science*,
400 320, 1456–1457.
- 401 [2] Malmshheimer RW, Bowyer JL, Fried JS et al. (2011) Managing forests because carbon matters:
402 Integrating energy, products, and land management policy. *Journal of Forestry* 109:S7–S51.
- 403 [3] Ter-Mikaelian M, McKechnie J, Colombo S, Chen J, MacLean H (2011) The carbon neutrality
404 assumption for forest bioenergy: a case study for northwestern Ontario. *Forest Chronicle* 87:644–
405 652.
- 406 [4] Helin T, Sokka L, Soimakallio S, Pingoud K, Pajula T (2013) Approaches for inclusion of forest
407 carbon cycle in life cycle assessment — a review. *Global Change Biology Bioenergy* 5:475–486.
- 408 [5] Mitchell SR, Harmon ME, O'Connell KEB (2012) Carbon debt and carbon sequestration parity in
409 forest bioenergy production. *Global Change Biology Bioenergy* 4(6):818–827.
- 410 [6] McKechnie J, Colombo S, MacLean HL (2014) Forest carbon accounting methods and the
411 consequences of forest bioenergy for national greenhouse gas emissions inventories.
412 *Environmental Science and Policy* 44:164–173.
- 413 [7] Routa J, Kellomäki S, Peltola H (2012) Impacts of intensive management and landscape structure
414 on timber and energy wood production and net CO₂ emissions from energy wood use of Norway
415 Spruce. *Bioenergy Research* 5:106–123.
- 416 [8] Garcia-Gonzalo J, Peltola H, Zubizarreta Gerendiain A, Kellomäki S (2007) Impacts of forest
417 landscape structure and management on timber production and carbon stocks in the boreal forest
418 ecosystem under changing climate. *Forest Ecology and Management* 241:243–257.
- 419 [9] Sathre R, Gustavsson L (2011) Time-dependent climate benefits of using forest residues to
420 substitute fossil fuels. *Biomass and Bioenergy* 35:2506–2516.

- 421 [10] Lundmark T, Bergh J, Hofer P et al. (2014) Potential roles of Swedish forestry in the context of
422 climate change mitigation. *Forests* 5(4):557–578.
- 423 [11] Kilpeläinen A, Torssonen P, Strandman H, Kellomäki S, Asikainen A, Peltola H (2016) Net
424 climate impacts of forest biomass production and utilization in managed boreal forests. *Global*
425 *Change Biology Bioenergy* 8:307–316.
- 426 [12] Cherubini F, Guest G and Strømman AH (2013) Bioenergy from forestry and changes in
427 atmospheric CO₂: reconciling single stand and landscape level approaches, *Journal of*
428 *Environmental Management* 129:292–301.
- 429 [13] Zanchi G, Pena N, Bird N (2012) Is woody bioenergy carbon neutral? A comparative assessment
430 of emissions from consumption of woody bioenergy and fossil fuel. *Global Change Biology*
431 *Bioenergy* 4:761–772.
- 432 [14] Levasseur A, Lesage P, Margni M, Samson R (2013) Biogenic carbon and temporary storage
433 addressed with dynamic life cycle assessment. *Journal of Industrial Ecology* 17:117–128.
- 434 [15] Routa J, Kellomäki S, Kilpeläinen A, Peltola H, Strandman H (2011a) Effects of forest
435 management on the carbon dioxide emissions of wood energy in integrated production of timber
436 and energy biomass. *Global Change Biology Bioenergy* 3:483–497.
- 437 [16] Routa J, Kellomäki S, Peltola H, Asikainen A (2011b) Impacts of thinning and fertilization on
438 timber and energy wood production in Norway spruce and Scots pine: scenario analyses based on
439 ecosystem model simulations. *Forestry* 84:159–175.
- 440 [17] Pyörälä P, Peltola H, Strandman H, Kilpeläinen A, Asikainen A, Jylhä K, Kellomäki S (2014)
441 Effects of management on economic profitability of forest biomass production and carbon
442 neutrality of bioenergy use in Norway spruce stands under the changing climate. *Bioenergy*
443 *Research* 7:279–294.
- 444 [18] Repo A, Tuovinen JP, Liski J (2015) Can we produce carbon and climate neutral forest
445 bioenergy? *Global Change Biology Bioenergy* 7:253–262.

- 446 [19] Sathre R, Gustavsson L (2012) Time-dependent radiative forcing effects of forest fertilization
447 and biomass substitution. *Biogeochemistry* 109:203–218.
- 448 [20] Haus S, Gustavsson L, Sathre R (2014) Climate mitigation comparison of woody biomass
449 systems with the inclusion of land-use in the reference fossil system. *Biomass and Bioenergy*
450 65:136–144.
- 451 [21] Torssonen P, Kilpeläinen A, Strandman H, Kellomäki S, Jylhä K, Asikainen A, Peltola H (2016)
452 Effects of climate change and management on net climate impacts of production and utilization of
453 energy biomass in Norway spruce with stable age-class distribution. *Global Change Biology*
454 *Bioenergy* 8:419–427.
- 455 [22] Nurmi J (1993) Small-sized trees above ground biomass heating value (Pienikokoisten puiden
456 maanpäällisen biomassan lämpöarvo). Helsinki. *Acta Forestalia Fennica* 236. 30p. (in Finnish).
- 457 [23] Nurmi J (1997) Heating values of mature trees. *Acta Forestalia Fennica* 256. 28p.
- 458 [24] Energy Statistics Yearbook (2011). Statistics Finland. Helsinki, Finland.
- 459 [25] Ramaswamy W, Boucher O, Haigh J et al. (2001) Radiative forcing of climate change. In:
460 Houghton JT et al. (eds) *Climate Change 2001: The Scientific Basis. Contribution of Working*
461 *Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change,*
462 *Cambridge University Press, Cambridge, UK and New York, NY. pp 349–416.*
- 463 [26] Foster P, Ramaswamy P, Artaxo T, Berntsen R, Betts DW et al. (2007) Changes in Atmospheric
464 Constituents and in Radiative Forcing. In: Solomon S, Qin D, Manning M, et al. (eds) *Climate*
465 *Change 2007: The Physical Science Basis. Contribution of Working Group I to the Fourth*
466 *Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University*
467 *Press, Cambridge, United Kingdom and New York, NY, USA. pp 130–234*
- 468 [27] Kellomäki S, Peltola H, Nuutinen T, Korhonen KT, Strandman H (2008) Sensitivity of managed
469 boreal forests in Finland to climate change, with implications for adaptive management.
470 *Philosophical Transactions of the Royal Society B*363:2341–2351.

- 471 [28] Kilpeläinen A, Alam A, Strandman H, Kellomäki S (2011) Life cycle assessment tool for
472 estimating net CO₂ exchange of forest production. *Global Change Biology Bioenergy* 3:461–471.
- 473 [29] Kellomäki S, Strandman H, Nuutinen T, Peltola H, Korhonen KT, Väisänen H (2005) Adaptation
474 of forest ecosystems, forest and forestry to climate change. FINADAPT. Working Paper 4. Finnish
475 Environment Institute Mimeographs 334. Helsinki.
- 476 [30] Mäkipää R, Karjalainen T, Pussinen A, Kukkola M, Kellomäki S, Mälkönen E (1998)
477 Applicability of a forest simulation model for estimating effects of nitrogen deposition on a forest
478 ecosystem: test of the validity of a gap-type model. *Forest Ecology and Management* 108:239–
479 250.
- 480 [31] Alam A, Kellomäki S, Kilpeläinen A, Strandman H (2013) Effects of stump extraction on the
481 carbon sequestration in Norway spruce forest ecosystems under varying thinning regimes with
482 implications for fossil fuel substitution. *Global Change Biology Bioenergy* 5:445–458.
- 483 [32] Venäläinen A, Tuomenvirta H, Pirinen P, Drebs A (2005) A Basic Finnish climate data set 1961–
484 2000— description and illustrations. *Reports of the Finnish Meteorological Institute* 5. 27p.
- 485 [33] Aalto J, Pirinen P, Heikkinen J, Venäläinen A (2012) Spatial interpolation of monthly climate
486 data for Finland: comparing the performance of kriging and generalized additive models.
487 *Theoretical and Applied Climatology* 112:99–111.
- 488 [34] Äijälä O, Koistinen A, Sved J, Vanhatalo K, Väisänen P (2014) Recommendations for Forest
489 Management in Finland. (in Finnish: Hyvän metsänhoidon suositukset – METSÄNHOITO),
490 Forestry Development Centre Tapio publications. 264 p. (in Finnish).
- 491 [35] Äijälä O, Kuusinen M, Koistinen A (2010) Recommendations for management and harvesting
492 of energy wood. (in Finnish: Hyvän metsänhoidon suositukset energiapuun korjuuseen ja
493 kasvatukseen), Forestry Development Centre Tapio publications. 32 p. (in Finnish).
- 494 [36] Schlamadinger B, Spitzer J, Kohlmaier GH, Ludeke M (1995) Carbon balance of bioenergy from
495 logging residues. *Biomass and Bioenergy* 8:221–234.

- 496 [37] Repo A, Tuomi M, Liski J (2011) Indirect carbon dioxide emissions from producing bioenergy
497 from forest harvest residues. *Global Change Biology Bioenergy* 3:107–115.
- 498 [38] Zetterberg L and Chen D (2014) The time aspect of bioenergy — climate impacts of solid
499 biofuels due to carbon dynamics. *Global Change Biology Bioenergy* 7:785–796.
- 500 [39] Ter-Mikaelian M, Colombo S, Chen J (2015) The burning question: Does forest bioenergy
501 reduce carbon emissions. A review of common misconceptions about forest carbon accounting.
502 *Journal of Forestry* 113:57–68.
- 503 [40] Gaudreault C and Miner R (2015) Temporal aspects in evaluating the greenhouse gas mitigation
504 benefits of using residues from forest products manufacturing facilities for energy production.
505 *Journal of Industrial Ecology* 19:994–1007.
- 506 [41] Mika AM, Keeton WS (2015) Net carbon fluxes at stand and landscape scales from wood
507 bioenergy harvests in the US Northeast. *Global Change Biology Bioenergy* 7:438–454.
- 508 [42] Gustavsson L, Haus S, Ortiz CA, Sathre R, Truong NL (2015) Climate effects of bioenergy from
509 forest residues in comparison to fossil energy. *Applied Energy* 138:36–50.
- 510 [43] Sedjo RA, Tian X (2012) An investigation of the carbon neutrality of wood bioenergy. *Journal*
511 *of Environmental Protection* 3:989–1000.
- 512 [44] Sathre R, Gustavsson L, Bergh J (2010) Primary energy and greenhouse gas implications of
513 increasing biomass production through forest fertilization. *Biomass and Bioenergy* 34:572–581.
- 514 [45] Mälkönen E (1976) Effect of whole-tree harvesting on soil fertility. *Silva Fennica* 10:157–164.
- 515 [46] Kuusinen M, Ilvesniemi H (eds) (2008) Energiapuun korjuun ympäristövaikutukset,
516 tutkimusraportti. Tapion ja Metlan julkaisuja, Helsinki, Finland, 74 p. (in Finnish).
- 517 [47] Jacobson S, Kukkola M, Mälkönen E, Tveite B (2000) Impact of whole-tree harvesting and
518 compensatory fertilization on growth of coniferous thinning stands. *Forest Ecology and Management*
519 129:41–51.

520 [48] Mäkipää R, Karjalainen T, Pussinen A, Kukkola M (1998) Effects of nitrogen fertilization on
521 carbon accumulation in boreal forests: model computations compared with the results of long-term
522 fertilization experiments. *Chemosphere* 36:1155–1160.

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552 **Figure and table captions**

553

554 **Fig 1.** Alternative initial stand age structures of the forest area producing energy biomass used in the
555 study. Right-skewed = forest area occupied initially by mostly young stands. Normal = forest area
556 occupied initially by mostly middle-aged stands. Left-skewed = forest area occupied initially by
557 mostly mature stands. Even = forest area consisting of stands each representing one age class.

558

559 **Fig 2.** Schematic figure of system boundaries of the study.

560

561 **Fig 3.** Cumulative mean energy biomass (logging residues, stumps, and roots) production from forest
562 areas with alternative initial stand age structures (see **Fig. 1**) and forest management scenarios over
563 the 80-year study period expressed in energy units. a = baseline with energy biomass harvesting
564 (BNR), b = nitrogen fertilization (BNRF), c = 20% higher stocking density (BNR20) and d = 20%
565 higher stocking density and nitrogen fertilization (BNR20F).

566

567 **Fig 4.** Net climate impact of energy biomass production and utilization (I_{NET}) (i.e. cumulative mean
568 difference in C_{net} between biosystem and fossil system) from forest areas with alternative initial stand
569 age structures (see **Fig. 1**) and forest management scenarios over the 80-year study period. a =
570 baseline with energy biomass harvesting (BNR), b = nitrogen fertilization (BNRF), c = 20% higher
571 stocking density (BNR20) and d = 20% higher stocking density and nitrogen fertilization (BNR20F).

572

573 **Fig 5.** Cumulative mean carbon neutrality of energy biomass production and utilization (CN) for
574 forest areas with alternative initial age structures (see **Fig. 1**) and forest management scenarios over
575 the 80-year study period. a = baseline with energy biomass harvesting (BNR), b = nitrogen
576 fertilization (BNRF), c = 20% higher stocking density (BNR20) and d = 20% higher stocking density
577 and nitrogen fertilization (BNR20F). Right-skewed = forest area occupied initially by mostly young
578 stands. Normal = forest area occupied initially by mostly middle-aged stands. Left-skewed = forest
579 area occupied initially by mostly mature stands. Even = forest area consisting of stands each
580 representing one age class.

581 **Fig 6.** Cumulative relative radiative forcing (CRF) of energy biomass production and utilization under
582 alternative initial age structures (See **Fig. 1**) and forest management scenarios over the 80-year study
583 period. a = Right-skewed, b = Normal, c = Left-skewed and d = Even. BNR = baseline with energy
584 biomass harvesting, BNRF = nitrogen fertilization, BNR20 = 20% higher stocking density and
585 BNR20F = 20% higher stocking density and nitrogen fertilization.

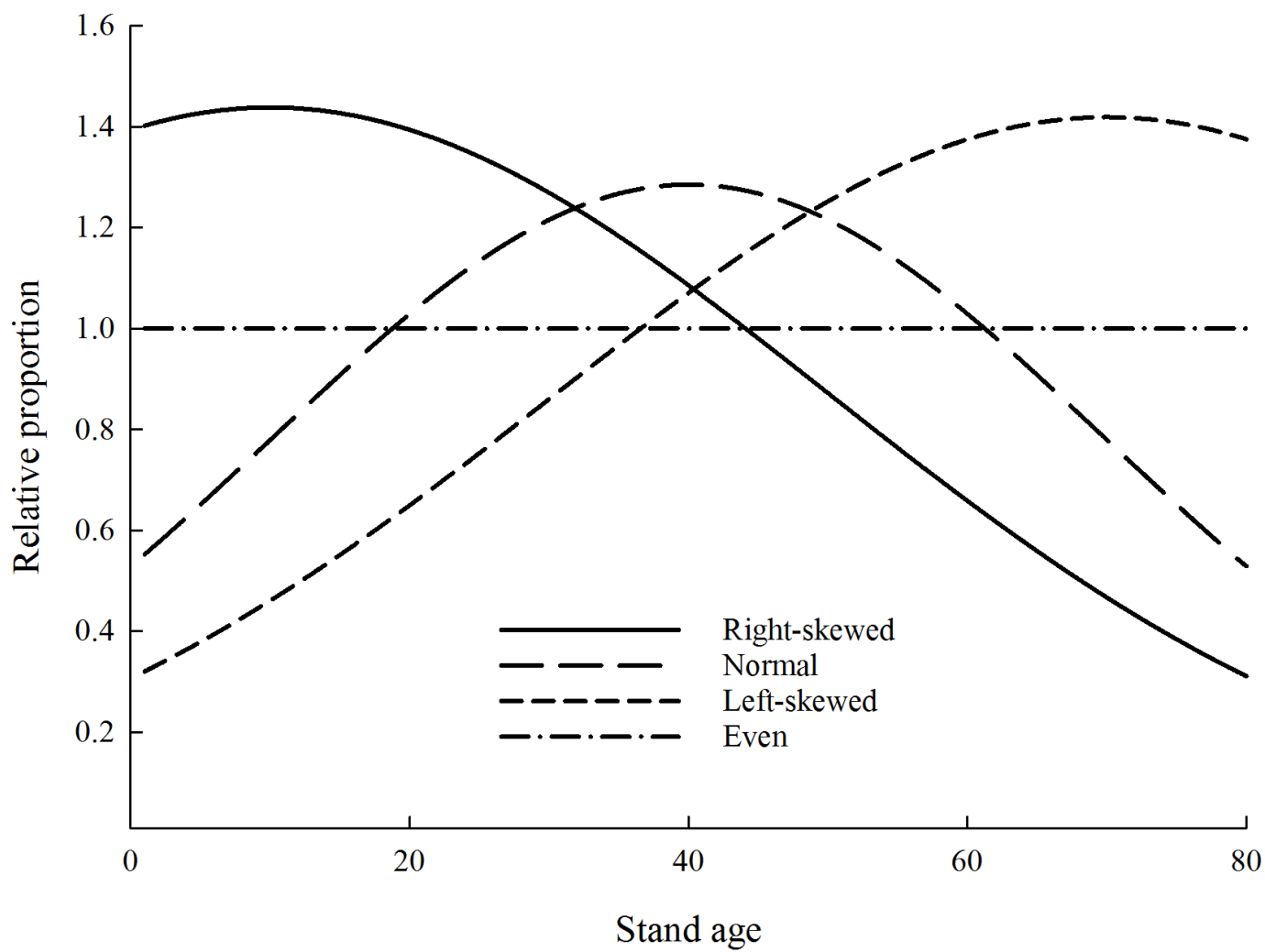
586 **Table 1.** Management and harvesting scenarios used in the study. BT = baseline (reference)
587 management without energy biomass harvesting, BNR = baseline management with energy biomass
588 harvesting, BNR20 = 20% higher stocking density compared to the baseline management, BNRF =
589 nitrogen fertilization, and BNR20F = 20% higher stocking density compared to the baseline
590 management and nitrogen fertilization.

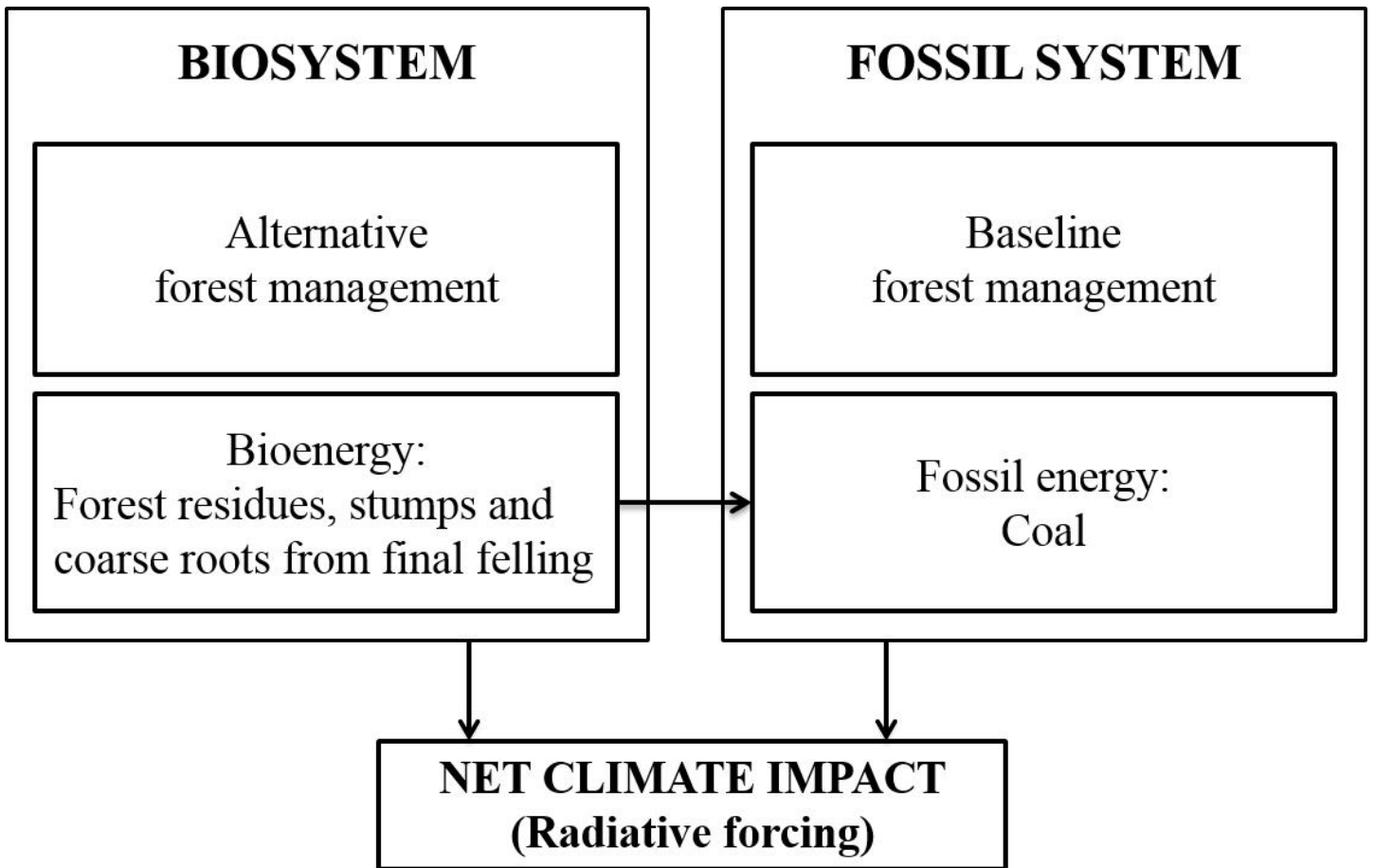
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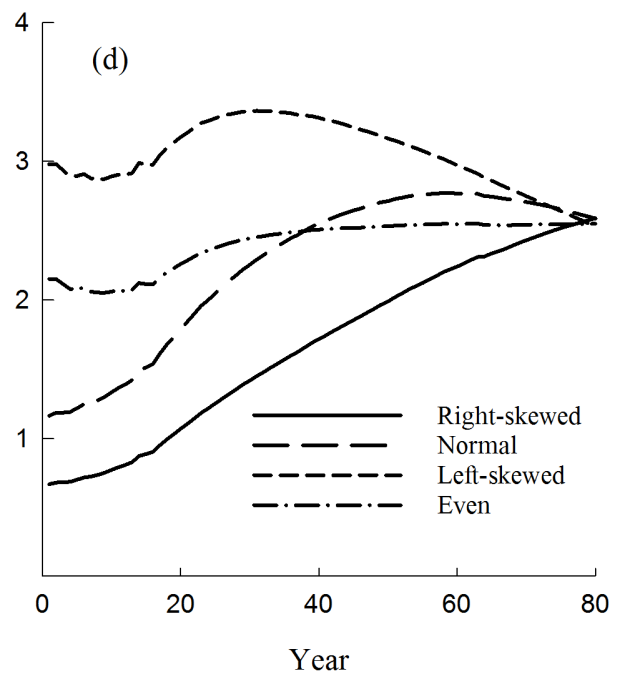
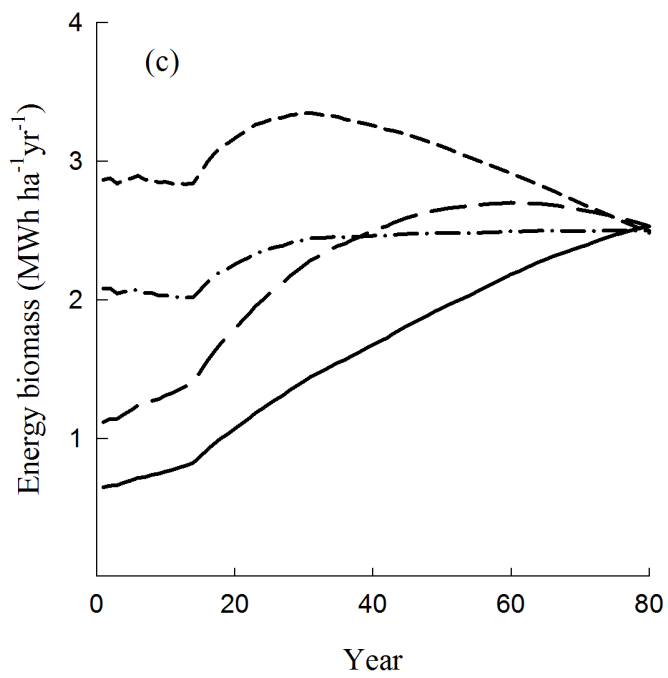
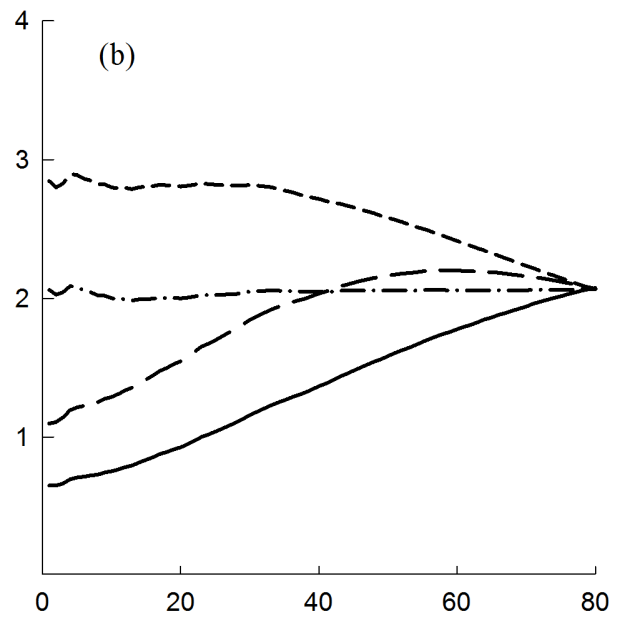
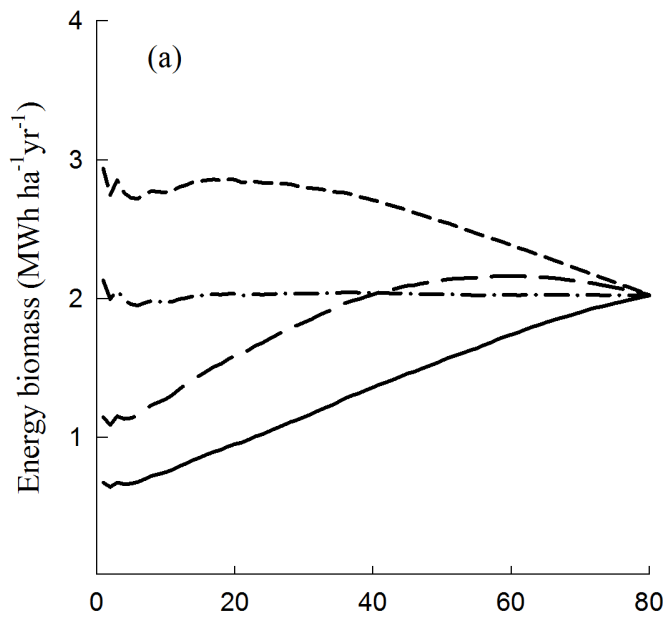
592 **Table 2.** Avoided CO₂ emissions of energy biomass production and utilization (kg CO₂ MWh⁻¹yr⁻¹)
593 under alternative forest management scenarios and initial age structures over the 80-year study period.
594 BNR = baseline management with energy biomass harvesting, BNR20 = 20% higher stocking density
595 compared to the baseline management, BNRF = nitrogen fertilization; and BNR20F = 20% higher
596 stocking density compared to the baseline management and nitrogen fertilization.

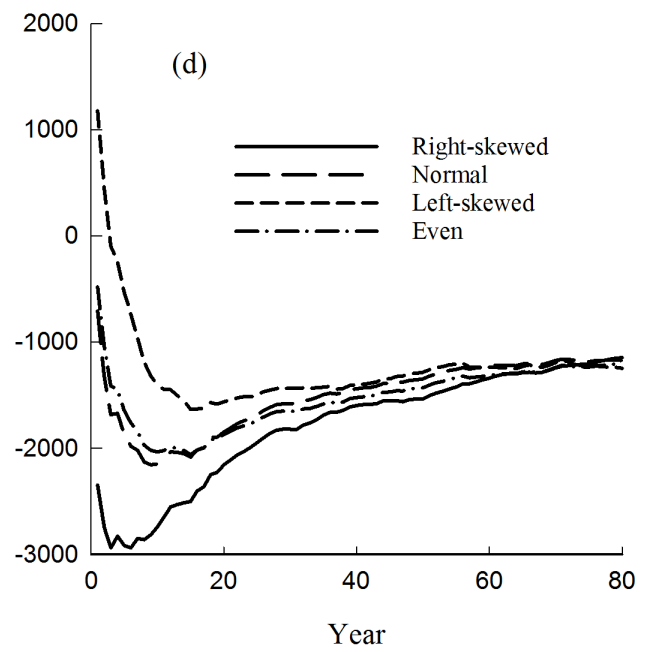
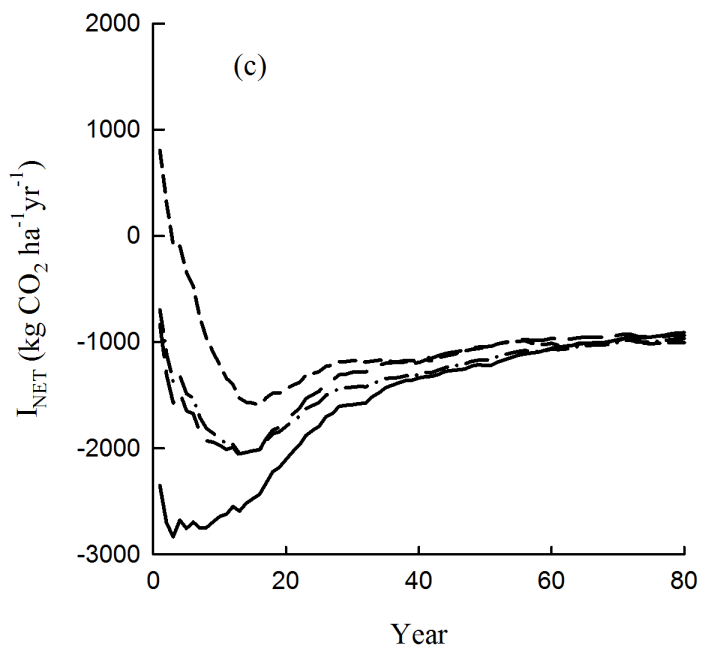
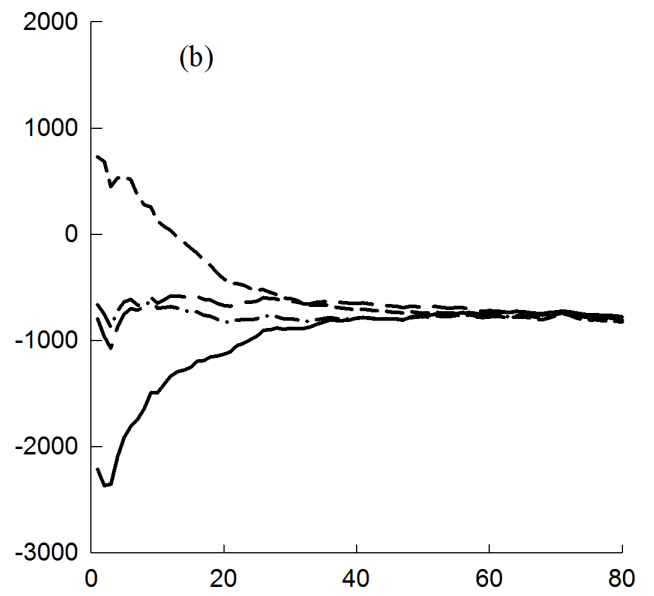
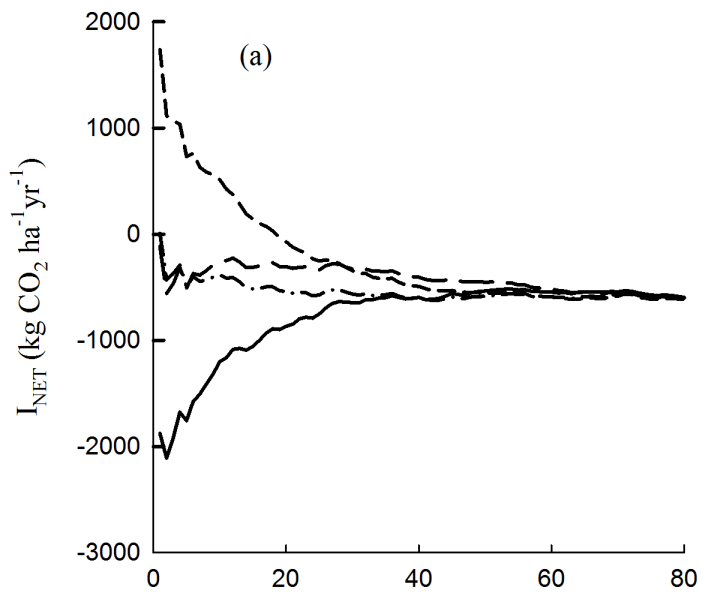
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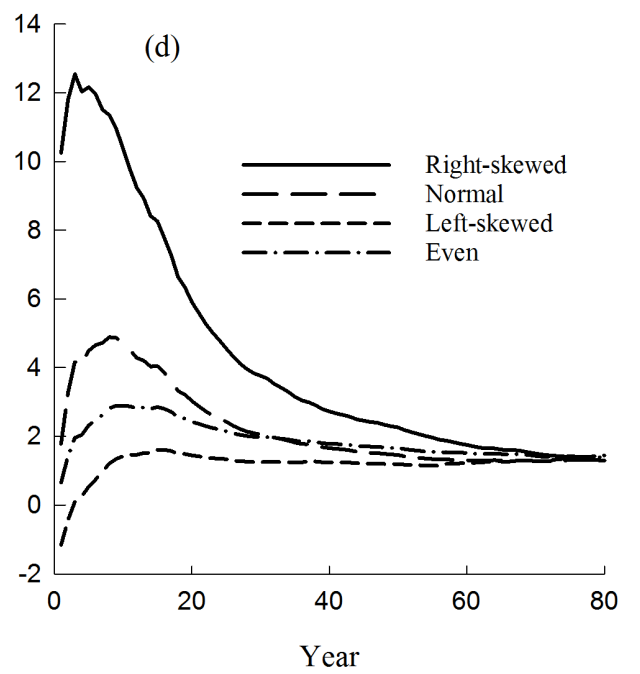
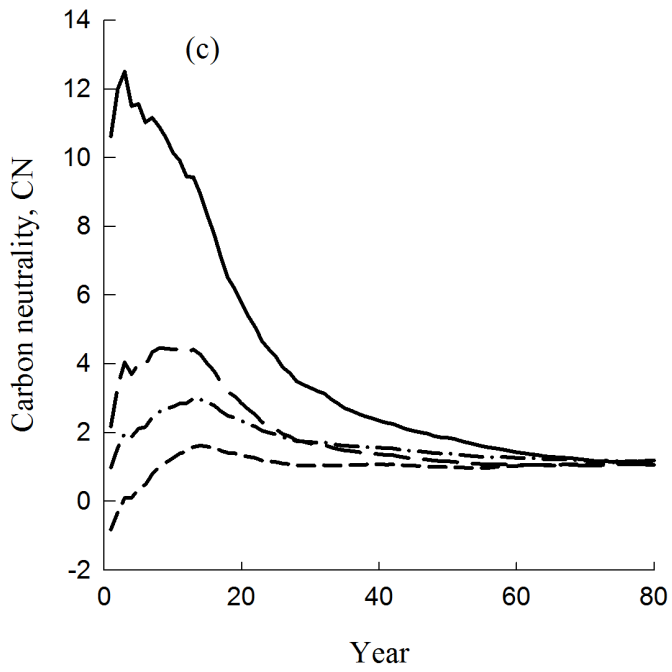
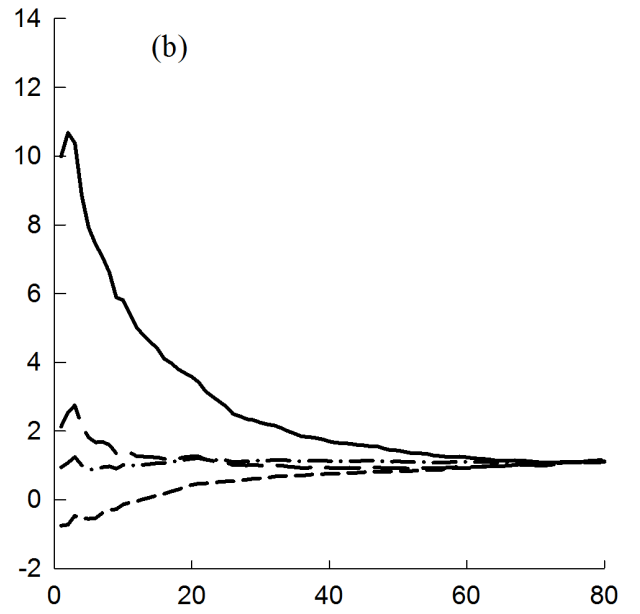
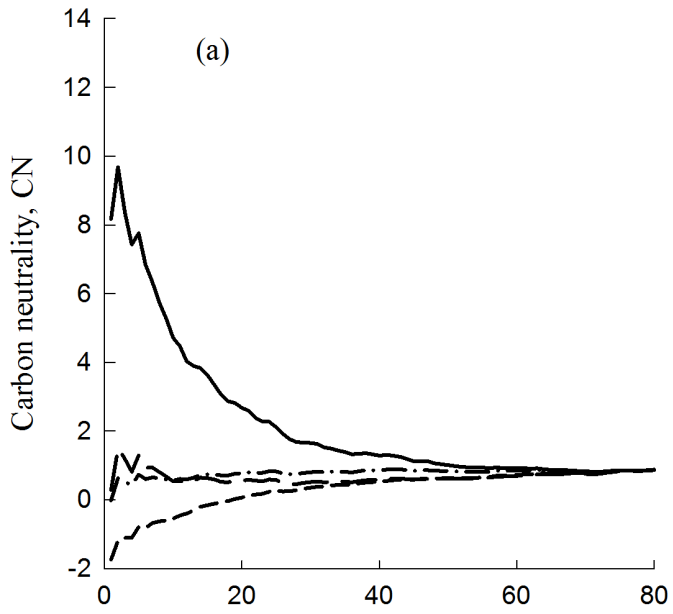
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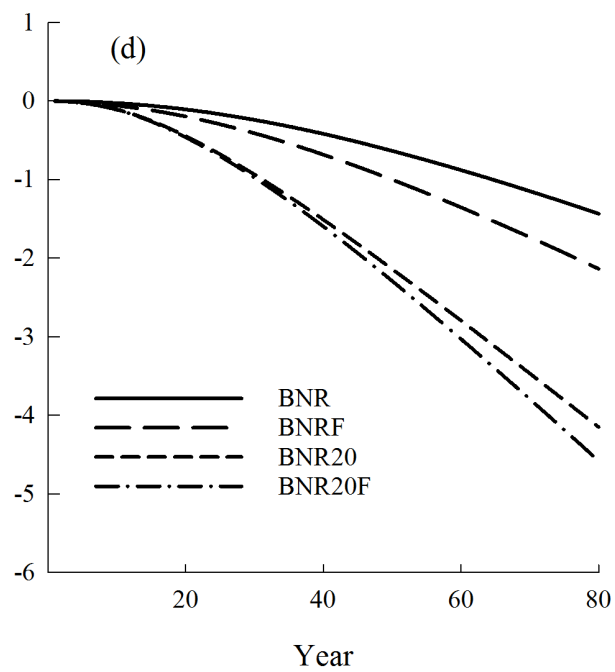
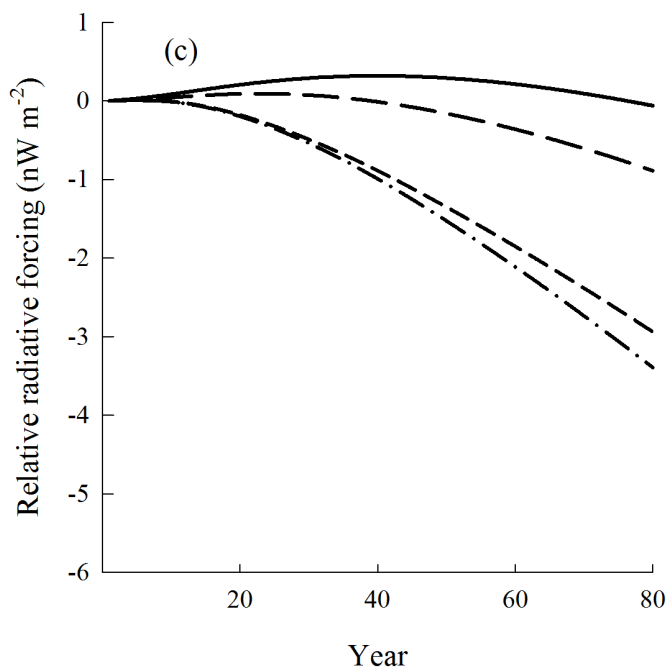
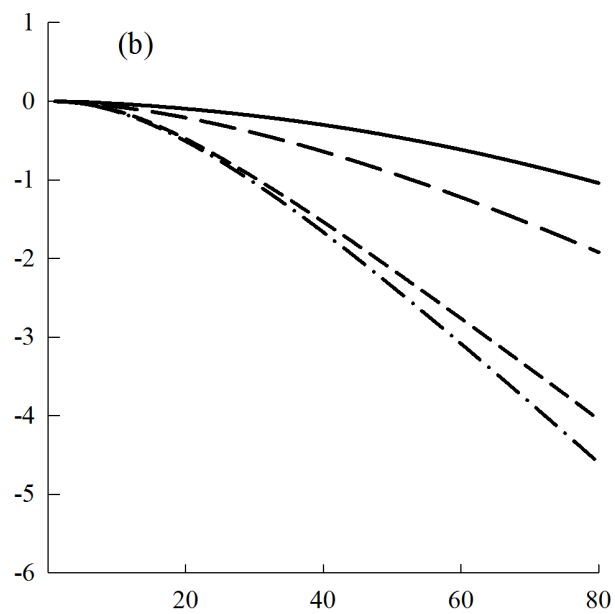
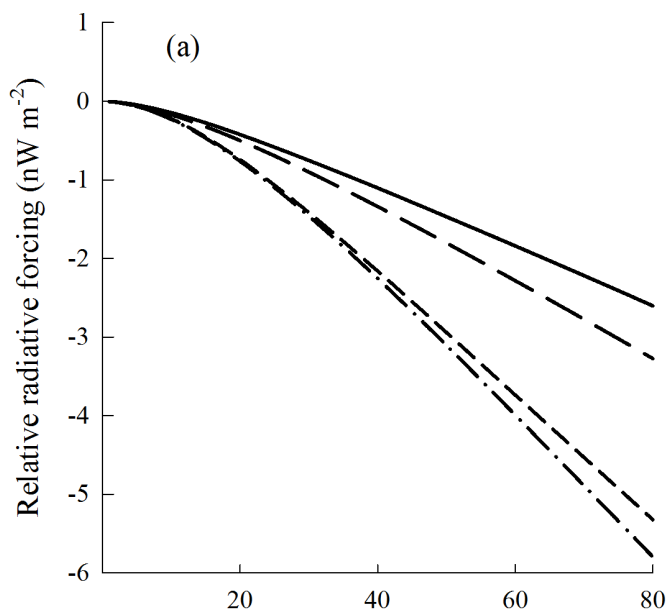


Table 1. Management and harvesting scenarios used in the study. BT = baseline (reference) management without energy biomass harvesting, BNR = baseline management with energy biomass harvesting, BNR20 = 20% higher stocking density compared to the baseline management, BNRF = nitrogen fertilization, and BNR20F = 20% higher stocking density compared to the baseline management and nitrogen fertilization.

Management scenario	Change in basal area thresholds, %	Fertilization, 2 x 150 kg N ha ⁻¹	Harvesting of energy biomass
BT(baseline)	-	-	-
BNR	-	-	Yes
BNR20	+ 20 %	-	Yes
BNRF	-	F	Yes
BNR20F	+ 20 %	F	Yes

Table 2. Avoided CO₂ emissions of energy biomass production and utilization (kg CO₂ MWh⁻¹yr⁻¹) under alternative forest management scenarios and initial age structures over the 80-year study period. BNR = baseline management with energy biomass harvesting, BNR20 = 20% higher stocking density compared to the baseline management, BNRF = nitrogen fertilization; and BNR20F = 20% higher stocking density compared to the baseline management and nitrogen fertilization.

Management scenario	Avoided CO ₂ emissions, kg CO ₂ MWh ⁻¹ yr ⁻¹			
	Initial age structure			
	Right-skewed	Normal	Left-skewed	Even
BNR	-413	-328	-410	-303
BNR20	-423	-311	-360	-283
BNRF	-424	-302	-329	-265
BNR20F	-388	-284	-343	-257