

2018

Is stump removal for bioenergy production effective in reducing pine weevil (*Hylobius abietis*) and *Hylastes* spp. breeding and feeding activities at regeneration sites?

Rahman, Abul

Elsevier BV

Tieteelliset aikakauslehtiartikkelit

© Elsevier B.V.

CC BY-NC-ND <https://creativecommons.org/licenses/by-nc-nd/4.0/>

<http://dx.doi.org/10.1016/j.foreco.2018.05.003>

<https://erepo.uef.fi/handle/123456789/7118>

Downloaded from University of Eastern Finland's eRepository

1 Is stump removal for bioenergy production effective in reducing pine weevil (*Hyllobius abietis*) and
2 *Hylastes* spp. breeding and feeding activities at regeneration sites?

3

4 Abul Rahman^a, Heli Viiri^b, Olli-Pekka Tikkanen^a

5

6 ^aUniversity of Eastern Finland, School of Forest Sciences, P.O. Box 111, FI-80101 Joensuu,
7 Finland.

8

9 ^bNatural Resources Institute Finland (Luke), Joensuu Research unit, P.O. Box 68, FI-80101
10 Joensuu, Finland.

11

12 *Corresponding Author. *E-mail address*: abul.rahman@uef.fi; Mobile: +358 440521229

13

14

15

16

17

18

19

20

21

22

23

24

25

26 **Abstract**

27

28 Stump harvesting can help in managing forest pests, improve site preparation, and provide a source
29 of bioenergy. However, stump removal does not remove all the roots from clear-cut areas. To
30 investigate whether stump removal helps to manage forest pests, the effect of stump removal and its
31 timing on the breeding and larval feeding activities of pine weevil (*Hylobius abietis*) and *Hylastes*
32 spp. was studied. In eastern Finland, 16 commercial regeneration sites dominated by Norway spruce
33 (*Picea abies*) (eight control areas, eight stump removal areas) were selected. Stumps were harvested
34 in 2011, within the year following logging in three of the stump removal sites (short delay
35 extraction), and in the second year after logging at five of the stump removal sites (long delay
36 extraction). Root samples were excavated from sites three years after logging to examine the
37 amount of roots, gnawing intensity, and density of larvae. In the control plots, gnawed root surface
38 areas were 24% and 50% greater than those in long delay and short delay stump removal sites,
39 respectively. After timing treatment, the estimated larval densities of both species were lower than
40 the estimated larval densities in the control sites. In conclusion, the timing of stump extraction may
41 partially regulate the breeding material and abundance of *Hylobius* and *Hylastes*. However, it is
42 probable that this effect is not strong enough to substantially limit the future damage on planted
43 seedlings.

44

45 **Key words:** Logging; Stump removal; Roots; Forest pests; Larvae.

46

47

48

49

50

51 **1. Introduction**

52 Tree stumps from forest regeneration areas are potentially an important source of raw material for
53 bioenergy production because stumps offer more biomass than logging residues (Egnell et al.,
54 2007). As well, stump harvesting may open new opportunities for managing forest pests and
55 diseases and improve quality in site preparation (Saarinen 2006). However, stump harvesting can
56 also adversely affect soil carbon stores, increase soil erosion and compaction, reduce soil nutrient
57 stocks, and cause valuable habitat loss for mosses, fungi, insects, etc. (Walmsley and Godbold,
58 2010).

59
60 Previous studies focused on the effects of stump harvesting on species dependent on dead wood
61 (Work et al., 2016; Victorsson and Jonsell, 2016; Shevlin et al., 2017). However, few studies have
62 focused on the effects of stump harvesting on pest populations, especially on the *Hylobius* genus
63 (Coleoptera: Curculionidae), one of the most common and abundant economic pests in conifer
64 seedling stands in Europe (Långström and Day, 2004). *Hylobius* breeds in conifer stumps and roots,
65 and hampers the restocking of regeneration sites. In addition, larvae of *Hylastes cunicularius* Er.,
66 and *Hylastes brunneus* Er., another potential but poorly studied pest group in conifer seedlings, and
67 longhorn beetles (Cerambycidae), often exist both in pine and spruce stumps and roots (Victorsson
68 and Jonsell, 2016).

69
70 In a fresh clear-cut area, stumps and logging residues emit volatile compounds (e.g. several
71 monoterpenes and ethanol) that attract potentially harmful insects to the site, including pine weevil
72 (*Hylobius abietis*), (Nordlander, 1987; Brattli et al., 1998) and *Hylastes* spp. (Joseph et al., 2001)
73 which reproduce in the stumps and roots of logged trees. Pine weevils lay their eggs in the soil and
74 bark of the roots (Nordlander et al., 1997) and *Hylastes* spp. also lay their eggs in recently clear-cut
75 stumps (Lindelöw et al., 1993). *Hylastes cunicularius* Er. breeds mainly in Norway spruce and

76 *Hylastes brunneus* Er. breeds mainly in Scots pine. After hatching, pine weevil larvae overwinter in
77 stumps, feeding under the bark of stumps and roots, and pupate in the following summer
78 (Nordenhem, 1989). New adult weevils emerge in autumn of the year following clear-cutting. In
79 this way, pine weevil breeding continues actively for a few successive years after clear-cutting has
80 occurred.

81
82 Pine weevils and *Hylastes* spp. preferentially feed on the thin bark of coniferous tree species
83 (Manlove et al., 1997; Leather et al., 1999; Löf et al., 2005; Wallertz et al., 2014). Pine weevils feed
84 on the roots and branches of mature trees and on the stems of seedlings. Both *Hylastes* species feed
85 on the roots of mature trees, and on the roots and at the stem base of seedlings, but just on the basis
86 of feeding marks it is impossible to separate the species. In boreal forest regeneration sites, pine
87 weevil feeding can cause the death of 60–80% of planted coniferous seedlings (Örlander and
88 Nilsson, 1999). Sustained pine weevil feeding on seedlings can last at least three consecutive years
89 (Långström, 1982). Most serious economic damage due to pine weevil feeding occurs at newly
90 planted coniferous regeneration sites where previous stands have been clear-cut coniferous forests
91 (Långström and Day, 2004).

92
93 In theory, the rapid harvesting of stumps and coarse logging residue after clear-cutting might
94 effectively reduce the amount of volatile compounds, which lure new adults to the clear-cut area. In
95 addition, it could also reduce the amount of suitable breeding material available and decrease the
96 subsequent larval population. Consequently, stump removal might reduce the feeding damage
97 caused by pine weevil and *Hylastes* spp. on planted seedlings. Thus, stump removal might function
98 as a silvicultural method in the integrated management of root-feeding pests. However, immediate
99 and total stump removal may not be possible in practical forestry management terms. In practice, in
100 stump harvesting, an excavator uproots the main tree root system, but many side roots and rotten

101 roots remain in the soil. Silvicultural instructions recommend to leave at least 25 stumps ha⁻¹ for
102 biodiversity and to prevent erosion (Koistinen, 2016). Moreover, stumps less than 20 cm in
103 diameter are often left due to the high cost of excavation (Kärhä, 2012).

104

105 The pine weevil has a strong ability to dig in the soil and lay eggs in small roots (Nordlander et al.,
106 1997). *Hylastes* spp. also can dig up to 100 cm in the soil to enter buried roots (Lindelöw, 1992).
107 Furthermore, if stump removal is delayed for a long time and done after arrival of pine weevils and
108 *Hylastes* spp. in clear-cut areas, then they have already succeeded in colonising the stump and root
109 system. This may compromise the potential pest control effect of stump removal. Therefore, it is
110 necessary to know how many roots are left for breeding substrate and how the timing of stump
111 removal in clear-cut areas contributes to the reproduction potential of *Hylobius abietis* and *Hylastes*
112 spp.

113

114 The concerns about the relationship of pine weevil and *Hylastes* spp. with stumps in clear-cut areas
115 are as follows: 1) both species feel attraction to stumps, and immigrate to clear-cut areas in early
116 summer; 2) weevils and *Hylastes* both breed in stumps and roots, and after completion of the larval
117 stage, emergence can take more than two years after immigration for young adult pine weevils, and
118 more than one year for *Hylastes* spp. in eastern Finland; 3) normally, stump removal will be carried
119 out in clear-cut areas after pest insect immigration. With this knowledge, theoretically, it can be
120 assumed that early stump removal might decrease the amount of breeding material and the
121 abundance of pine weevil and *Hylastes* spp. in the regeneration site. To determine how stump
122 removal and its timing affect the breeding and abundance of pine weevils and *Hylastes* species, we
123 studied the effects of stump removal on the amount of coniferous root material remaining in clear-
124 cut areas available for *Hylobius* and *Hylastes*. We also tested short delay (within a year of clear-cut)
125 versus long delay (the year following clear-cut) stump removal on the populations of *H. abietis* and

126 *Hylastes* spp. by measuring their larval feeding intensities on roots. *Hylastes* spp. and pine weevil
127 feeding intensities in the remaining roots were compared between control sites (with no stump
128 removal) and sites with stump removal.

129

130 **2. Material and methods**

131 *2.1. Study sites and experimental design*

132 In this study, 10 and 6 regeneration sites, logged in 2009 and 2010, respectively were selected in
133 eastern Finland (Table 1). The sites were dominated by Norway spruce (*Picea abies* L.). Typically,
134 in January and July, the mean temperature is -16°C and +17°C, respectively, in North Karelia (data
135 from Finnish Meteorological Institute). In this experiment, sites were clear-cut in the previous
136 winter season and fresh stumps were available for insect colonisation in the following spring. The
137 experiment had a paired-site design: the 16 sites were paired, so that each pair consisted of a control
138 site (stumps left intact) and a stump removal site. Control and stump removal sites were paired
139 based on approximately equal volumes of standing stock before final logging. Because of the time
140 lag between logging and stump extraction, stump extraction sites were classified in either the short
141 delay or long delay category. The five pairs logged in 2009 (stump extracted 2010) were those with
142 a long delay and the three pairs logged in 2010 (stump extracted 2010) were those with a short
143 delay between the logging and stump harvest.

144

145 Stump removals were performed on all experimental sites in 2010 according to the normal
146 commercial practices and instructions approved by good silvicultural practise for energy wood
147 harvest (Koistinen et al., 2016). According to the instructions, more than 25 stumps of various tree
148 species ha⁻¹ (over 15 cm in diameter), all rotten stumps, and stumps less than 20 cm in diameter,
149 should be left. Each site had been mounded by an excavator and planted with Norway spruce
150 seedlings according to normal local forestry practices. All study sites were located on mineral soil.

151 The dominant late-successional forest floor species was *Vaccinium myrtillus* L. (mesic, *Myrtillus*
152 forest-type) for all sites except the Kermansalo and Jalaslampi sites, which had an herb-rich forest
153 type (*Oxalis-Myrtillus* type) (Cajander, 1949).

154

155 In the middle of each regeneration site, 20 sample plots were established. Sample plots were 1 m²
156 each and arranged along two lines, each line containing 10 sample plots. Sample plots had a
157 minimum distance of five meters between them. If the central point of the root extraction plot and
158 control site contained a big stone or stump, the sample plot was moved forward.

159

160 2.2. Root sampling

161 All roots of each sample plot were excavated manually in the autumn of 2012 (logged in 2009) and
162 2013 (logged in 2010). Litter, branches, and visible deciduous roots were removed from the sample
163 plot before excavation. Roots clearly identified in the field as belonging to a deciduous tree species
164 were ignored. All excavated conifer roots were put in plastic bags in the field, frozen in the
165 laboratory, and later identified in the laboratory. If pieces of bark from the roots fell off during
166 digging, the bark samples were put in the same bag as the root sample and was also examined later
167 for traces of insects, but generally, root decomposition had not progressed to the point that it would
168 have been disturbed by root sampling. Pits in the sampling plots were dug to the depth at which no
169 roots could be found. Roots that extended outside of the sample plot were cut exactly from the
170 border line of the plot with secateurs or a saw. In this study, sampled roots diameter was short delay
171 6.5 ± 4 cm (mean \pm SD), long delay 6.6 ± 4.6 cm, and control 9.8 ± 12.8 cm and the length short
172 delay 22 ± 11 cm, long delay 29 ± 15 cm, and control 32 ± 14 cm. However, in control plots there
173 were a few roots with root neck and part of the stump attached. The number of very large roots in
174 control plots were so small that it has no effect on results.

175

176 In the laboratory, the length and diameter of the root samples were measured from both ends of
177 each root. The root surface area was calculated using the formula $2[\pi r^2] + [2 \pi r] \times h \text{ cm}^2$, where r
178 is the mean radius of the sample root and h is the length of the sample root. The amount of feeding
179 by both *Hylobius* and *Hylastes* species was estimated for each root. Both species mine larval tunnels
180 in the phloem of the roots. *Hylobius* larval tunnels are about 5 mm wide. *H. cunicularius* and *H.*
181 *brunneus* tunnels cannot be separated from each other, they are too narrow, and occur mostly in
182 wood; thus the two species were grouped in this study. All discovered adult insects, larvae, and
183 pupal chambers were counted and identified. Gnawed root surface areas were calculated by [surface
184 area of each root \times 100]/gnawed (%) for each root. The total root surface area and gnawed root
185 surface areas were calculated for each plot. The fungal coverage of root samples was estimated.

186

187 In total, 8 891 coniferous root samples were examined. In this experiment, it was estimated by
188 assuming that one gnawed root indicated the presence of one larva. Total gnawed roots were
189 counted for each sample plot. To obtain the larval density ha^{-1} , we calculated the average number of
190 larvae present in the control, short delay stump removal, and long delay stump removal plots. Each
191 treatment averages were multiplied by 10 000 to obtain the per-hectare value of larval density.

192

193 2.3. Statistical analysis

194 Before statistical analyses, dependent variables were transformed by $\log 10+1$ to reduce non-
195 normality. We developed a mixed linear model in the following form:

$$196 \text{Log } Y_{ij} = \log \beta_1 + \log \beta_2 T_{ij} + \log \beta_3 \delta_{ij} + \log \beta_4 \theta_{ij} + \varepsilon_{ij}.$$

197 In the model, β_1 is constant, and β_2 , β_3 , and β_4 are the coefficients of the corresponding variables. T_{ij}
198 is the treatment as $T \in (\text{control [no stump removal]}, \text{SR [stump removal]})$, δ_{ij} is the treatment \times
199 time (short, long), θ_{ij} is the stand volume, i is the site, j is the paired sites (control, stump removal),
200 and ε_{ij} is the error in the model.

201

202 In the model, the total root surface area, gnawed root surface area, and larvae density were
203 dependent variables, the treatment and stump removal time difference were set as factors, and the
204 previous stand volume was considered as a covariate. Further, the treatment, time, treatment \times time
205 interaction, and stand volume were set as fixed effects. Sites and paired sites were set as random
206 effects. Significance levels were set at $p < 0.05$. In SPSS 17 statistical software, analyses of
207 variance were performed using a general linear mixed model.

208

209 **3. Results**

210 *3.1. Stump removal and availability of feeding resources*

211

212 In control sites, the mean root diameter was 34% larger than that of the stump removal sites (Table
213 2). There were significant variations in the root surface area between the control and stump removal
214 sites, and between the timing of the stump harvest (Table 3). Control sites had, respectively, 44%
215 and 64% greater mean root surface areas than the long delay (following year of clear-cut) and short
216 delay (within year of clear-cut) stump removal sites. Short delay stump removal sites had less root
217 surface area remaining than long delay removal sites, and control sites also showed the same pattern
218 (Fig. 1). The logging volume of the previous stand had no significant influence on the remaining
219 root surface area (Table 3), because sites were paired with corresponding volumes of the previous
220 stand as part of the experimental design.

221

222 *3.2. Effect of stump removal on number of larvae*

223 The total number of pine weevil pupal chambers found in the roots was 226, which is 2.5 % of the
224 total roots collected from all study sites (Fig 2.). Pupal chambers were more frequent in the control
225 sites (Fig. 2). Most pine weevil larvae (187) were also found in control sites, and only one pine

226 weevil larvae was found from a stump removal site (Fig. 2). In the stump removal sites, we
227 estimated the pine weevil larval density to be 78 780 and 35 700 larvae ha⁻¹ on the basis of the root
228 feedings, which were 6% and 48% lower than the estimated larval densities in the control sites in
229 long delay and short delay stump removal respectively (Fig 3). Stump extraction had a significant
230 effect on the larval density (Table 4). There was also a significant difference in pine weevil larval
231 density between treatment and time interaction (Table 4). Stump removal sites also had 21% fewer
232 *Hylastes* spp. feeding than control sites, and stump removal significantly reduced *Hylastes* spp.
233 feeding (Fig. 4) (Table 4).

234

235 3.3. Effect of stump removal on feeding

236 In stump removal sites, the gnawed root surface areas by *H. abietis* and *Hylastes* were 24% and
237 50% lower in the long delay and short delay plots than those in the control plots, respectively, and
238 the differences were significant (Fig. 5) (Table 5). Gnawed root surface areas were greater in the
239 long delay than short delay stump removal sites. In general, the gnawed root surface area in the
240 control areas was 34% larger than that in the stump removal sites.

241

242 In control sites, pine weevils gnawed 17% of roots; in stump harvesting sites, pine weevils gnawed
243 11% of roots. Pine weevils gnawed 38% more root surface area in control sites compared to stump
244 removal sites (Fig. 6). In addition, the area gnawed by *Hylastes* species was 26% larger in the
245 control sites than in the stump removal sites. Pine weevil gnawed more root surface area compared
246 to *Hylastes* spp. in both the control and stump removal sites (Fig. 6). In the mixed model, both
247 species each showed significant differences based on stump extraction treatments, but did not
248 exhibit significant differences based on time, treatment × time interaction, and logging volume
249 (Table 6). Traces of fungal growth were found in 3% of roots.

250

251 **4. Discussion and conclusion**

252 Our study showed that removal of stumps from clear-cut areas reduced the amount of suitable
253 breeding material with significant effect on the size of both pest populations in regeneration sites.
254 However, in the current stump harvesting practice it is not possible to eradicate root dwelling pests
255 completely from a clear-cut area. It was established in our study for the first time, that
256 approximately 35 700–78 780 pine weevil and 33 250–67 660 *Hylastes* spp. larvae ha⁻¹, or more,
257 were present in the remaining roots after stump removal. Because we used very strict criteria, that a
258 root with signs of feeding indicates the presence of only one larva, these values more likely are
259 closer to the minimum population densities than the maximum population densities.

260

261 Roots infected by certain fungi can be totally devoid of insect larvae (Skrzecz, 2017). Here,
262 however, only a very small proportion of roots were infected by fungi, and it seemed that fungal
263 decomposition had not yet destroyed the traces of larval tunnels or pupal chambers. Thus, the
264 timing of the root sampling fitted well to the local development phase of the pine weevils. The
265 number of pine weevil larvae found supported this, and gave justification to the estimation method
266 for determining pine weevil larval density.

267

268 Previously, Moore et al. (2004) have estimated that in Scotland, in areas where stumps were left
269 intact, the larval population of *H. abietis* was between 46 000–170 000 larvae ha⁻¹. Our estimate of
270 the larval population size on the control sites fits well within these limits. To our knowledge, there
271 are no other estimates of the larval population size of *H. abietis*, and no studies on the effect of
272 stump removal on it, excluding one older study performed in Sweden and published in Swedish (see
273 Långström and Day, 2004). Further, our results support the earlier speculation that the role of
274 *Hylastes* as a pest in conifer regeneration sites might be underestimated (Lindelöw, 1992).
275 Recently, Nordlander et al. (2017) provided that clear-cut sites on seedling mortality caused by

276 *Hylastes* spp. (mean 4 %) and *Hylobius* (mean 29 %) after two seasons. It is also important to
277 mention that less research has been carried out on *Hylastes* spp. compared to that on *Hylobius*
278 *abietis*. After all, it seems that the population density of *Hylastes* spp. is not much smaller than that
279 of *H. abietis*.

280

281 We found that the reducing effect of stump removal was more effective when stump extraction was
282 carried out with only a short delay after logging than if it was delayed for a longer time (extraction
283 in following year after clear-cut). A long delay between logging and stump extraction leads to only
284 a minor decrease in the number of roots with signs of feeding by larvae. In stands with a long delay
285 extraction treatment, higher amounts of stumps and roots were available for a longer time, which
286 allows a longer window of opportunity for colonisation and the spread of larvae into the root
287 systems of logged trees than in stands with a short delay stump extraction treatment. A short delay
288 between logging and the removal of fresh stumps may reduce the amount of attractive volatiles
289 emitted from the stand, and shorten the period when root dwelling pests are attracted to a clear-cut
290 area for breeding.

291

292 In addition to the timing of stump extraction, differences in the weather conditions between years
293 and sites might have some unknown effects on these results. The summer 2010 was extremely hot,
294 with an all-time summer temperature record of + 37°C in eastern Finland. Moreover, the sites
295 logged in winter 2010 (short delay sites) had a slightly more southern location than the sites logged
296 in 2009 (long delay sites). However, the weather-related factors probably have no major effect on
297 the results. The hot summer in 2010 affected all sites, and the cumulative temperature sums of the
298 three subsequent summers was almost the same between the periods 2009–2012 and 2010–2013; it
299 was only 72 degree days higher in the first period than in the second (data from Joensuu Airport
300 weather station, in the middle of the study areas; Finnish Meteorological Institute).

301

302 Although the reduction in size of estimated larval populations was significant in the short delay
303 stump extraction treatment, the size of estimated larval population still remained at a very high
304 level. These remaining larvae will mature and likely cause seedling damage during their emerging
305 period from the roots. Accordingly, parent pine weevils and new-born weevils may remain in high
306 numbers for some consecutive years following clear-cutting of a particular site (von Sydow 1997;
307 Örlander et al., 1997). Pine weevils have the ability to emerge from even four-year-old roots
308 (Nordenhem, 1989). Additionally, *Hylastes* species can remain for four to six years in clear-cut
309 areas and cause seedling damage (Lindelöw, 1992). However, if the stump harvesting operation can
310 be carried out before migration of adult pests into the clear-cut area, then pine weevil and *Hylastes*
311 spp. abundance might be minimised. To do this, forest managers must make proper decisions about
312 stump harvesting time.

313

314 Our study showed that there are a substantial number of *Hylastes* spp. present both in stump
315 extracted and control sites, but it seems that the stump removal had a more pronounced effect on the
316 pine weevil larvae populations than on populations of *Hylastes*. In control sites, *Hylastes* spp.
317 exhibited less root-feeding activity than pine weevils. However, after stump removal treatment, pine
318 weevil root feeding was reduced more than feeding by *Hylastes* species. This may be because
319 *Hylastes* utilizes recently died and dying roots for breeding (Ehnström and Axelsson, 2002) and can
320 be more abundant in mature forests than pine weevil (Heikkala, 2016). Moreover, pine weevil may
321 exhibit life cycle and behaviour that is strongly related to fresh clear cuts, and, therefore, stump
322 extraction reduced significantly pine weevil's attraction to harvested sites of this study. Whereas
323 *Hylastes* species were less affected because it was already present in logged stands.

324 Although it could be assumed that after clear-cut, *Hylastes* spp. and pine weevil larvae compete
325 food resources, we found that there were plenty of roots available for larval feeding for both

326 species. Large amounts of roots without signs of gnawing indicated that the competition pressure on
327 root dwelling larval populations was generally at a low level.

328

329 According to our results, stump removal reduces the amount of resources available for root feeding
330 pests and the size of the larval population. However, the critical question is that is this reduction
331 enough to diminish the size of the adult population, which causes the seedling damage, so that the
332 risk of future damage to seedlings would be considerably lower. In our study, it was estimated that
333 approximately 89 090 to 100 840 pine weevil larvae and 77 280 to 81 850 *Hylastes* spp. larvae were
334 still present in the clear-cut area. When long delay and short delay treatments were applied, then 78
335 780 and 35 700 pine weevil larvae, and 67 660 and 33 250 *Hylastes* spp. larvae survived,
336 respectively. According to Moore et al. (2004), 40–80% of pine weevil larvae survived to become
337 adults in traditional regeneration sites. If we follow their lowest survival rate estimation (40%), then
338 in our stump extracted site the number of emerging pine weevil adults ha⁻¹ will be 31 512 in long
339 delay and 14 280 in short delay sites. In addition, the number of emerging adults of *Hylastes* spp.
340 will be 27 064 (long delay sites) and 13 300 (short delay sites) ha⁻¹.

341

342 We have previously found that traditional regeneration sites, there was more seedling damage
343 caused by pine weevil than in sites with the stumps extracted (Rahman et al., 2015). In fresh clear-
344 cut areas, it has been estimated that 14 000 adult immigrant weevils ha⁻¹ can cause damage to 82%
345 of seedlings (Nordlander et al., 2003). Based on this finding and our calculations above, it is not
346 justified to conclude that the reduction in population size of weevils resulting from stump removal
347 is sufficient significantly to reduce the damage level to coniferous seedling stocks. Therefore, it
348 appears that stump extraction is not a very effective method to control damage caused by *Hylobius*
349 and *Hylastes*. Only if the stump extraction is performed without delay will there be a good

350 possibility of reducing the population size, and enhancing the impact of other first-line control
351 methods.

352

353 After all, in addition to pest damage control, stump removal has a multitude of other silvicultural
354 and ecological effects, which must be considered. Stump harvesting reduces the amount of ground
355 vegetation such as cowberry and bilberry (Andersson et al., 2016), and partially reduces moss cover
356 (Hyvönen et al., 2016). Several studies have suggested that intensive stump harvesting is a threat to
357 forest biodiversity, especially for species dependent on wood (Jonsell and Schroeder, 2014;
358 Victorsson and Jonsell, 2016; Shevlin et al., 2016). Potentially, stump wood for bioenergy is
359 beneficial to mitigating CO₂ emissions (Ortiz et al., 2016) but there are also reports claiming that
360 the opposite is the case (Mäkipää et al., 2015). In addition, in the field of forest protection, stump
361 harvesting has the potential to reduce infections of the root rot fungus *Heterobasidion* by 20–72%
362 in the next generation trees (Cleary et al., 2013). Theoretically, stump harvesting can reduce pine
363 weevil damage and lower the rate of root rot fungus infection. However, for practical applications
364 and to balance the expense of stump harvesting, forest managers have to make proper decisions and
365 identify suitable sites for stump harvesting for better forest management.

366

367 In the existing forest management system, it is not possible to reduce the breeding material and size
368 of weevil populations sufficiently with stump harvesting. It seems that by adjusting the time lag
369 between logging and stump extraction, it is possible to some extent to regulate the amount and
370 quality of remaining roots and thus reduce breeding material for pine weevil. The total elimination
371 of root material suitable for the breeding of root dwelling pests is technically difficult and not
372 possible with current stump extraction methods. Total removal of the roots would be highly
373 expensive and ecologically problematic. Instead of this unrealistic method, the rapid removal of cut
374 stumps might be applied to support other control methods for root-feeding pests. It is already well

375 established, and it has also been recently suggested that seedlings planted on mineral soil are
376 relatively safe from weevil damage (Luoranen et al., 2017). Additionally, as the seedlings grow,
377 having large stem diameters reduces pine weevil feeding damage (Nordlander et al., 2011).
378 Recently, Viiri and Luoranen (2017) suggested that deep planted seedlings can reduce pine weevil
379 feeding damage. If stump removal remains the forest manager's methodology in the future, it
380 should be studied whether it is possible to reduce weevil damage more effectively by combining
381 fast stump removal and choosing the planting spots with optimal conditions for seedling survival
382 and growth.

383

384 **Acknowledgements**

385 We thank Forest and Park Service for providing the study sites, Heimo Tynkkynen for root
386 sampling, and Seija Repo for root measurements. We also thank Lauri Mehtätalo for advising on
387 analytical procedures for our data. The study was financed by the Finnish Forest Research Institute
388 (Project 3565 "Influence of stump removal on forest pests") as part of the Forest Energy 2020
389 Research Programme and Finnish Cultural Foundation grants (Grant No. 22356). The manuscript
390 language was checked by Editage.com. We also thank the UEF Doctoral Programme in Forests and
391 Bioresources for their financial and logistic support to pursue this study.

392

393 **Figure captions**

394

395 **Fig 1.** Estimated marginal means (\pm S.E.) of the available surface area of roots in stump removal and
396 control sites with long delay and short delay time lag difference between logging and stump
397 extraction.

398

399 **Fig. 2.** Total identified and counted (\pm S.E.) pupal chambers, larvae of *H. abietis* during laboratory
400 assessment of the roots collected from stump removal and control sites.

401

402 **Fig 3.** Estimated marginal means (\pm S.E.) of pine weevil larval density in stump removal and control
403 sites with long delay and short delay time lag difference between logging and stump extraction.

404

405 **Fig 4.** Estimated marginal means (\pm S.E.) of *Hylastes* spp. larval density in stump removal and
406 control sites with long and short delay time lag difference between logging and stump extraction.

407

408 **Fig 5.** Estimated marginal means (\pm S.E.) of the gnawed area of roots (Long delay and short delay
409 time lag difference between logging and stump extraction x Treatment interaction) by *Hylastes* spp.
410 and *Hylobius abietis*.

411

412 **Fig 6.** Estimated marginal means (\pm S.E.) of the gnawed area of roots in stump removal and control
413 sites independently by *Hylastes* spp. and *Hylobius abietis*.

414

415

416

417

418

419

420

421 **References**

422

423 Andersson, J., Dynesius, M., and Hjältén, J., 2017. Short-term response to stump harvesting by the
424 ground flora in boreal clear-cuts. *Scand. J. For Res.* 32, 239–245.

425

426 Brattli, J. G., Andersen, J., and Nilssen, A.C., 1998. Primary attraction and host tree selection
427 indeciduous and conifer living Coleoptera: Scolytidae, Curculionidae, Cerambycidae and
428 Lymexylidae. *J. Appl. Entomol.* 122, 345–352.

429

430 Cajander A. K., 1949. Forest types and their significance. *Acta Forestalia Fennica.* 56(5),1–36.
431 doi.org/10.14214/aff.7396

432

433 Cleary, M.R., Arhipova, N., Morrison, D.J., Thomsen, I.M., Sturrock, R.N., Vasaitis, R., Gaitnieks,
434 T., and Stenlid, J., 2013. Stump removal to control root disease in Canada and Scandinavia: A
435 synthesis of results from long-term trials. *Fores. Ecol. Manage.* 290, 5–14.

436

437 Egnell, G., Hyvönen, R., Högbom, L., Johansson, T., Lundmark, T., Olsson, B., Ring, E., and von
438 Sydow, F., 2007. Miljökonsekvenser av stubbskörd – en sammanställning av kunskap och
439 kunskapsbehov. *Energimyndigheten Rapport ER 2007: 40* [In Swedish with English summary]

440

441 Ehnström, B and Axelsson, R., 2002. Insektsnag I bark och ved. *Artdatabanken.* SLU.

442

443 Heikkala, O., 2016. Emulation of natural disturbances and the maintenance of biodiversity in
444 managed boreal forests: the effects of prescribed fire and retention forestry on insect assemblages.

445 *Dissertationes Forestales.* 222, 1-46.

446

447 Hyvönen, R., Kaarakka, L., Leppälampi-Kujansuu, J., Olsson, B.A., Palviainen, M., Vegerfors, B.,
448 and Helmisaari H.-S., 2016. Effects of stump harvesting on soil C and N stores and vegetation 8–13
449 years after clear-cutting. *Fores. Ecol. Manage.* 371, 23–32.

450

451 Jonsell, M., and Schroeder, M., 2014. Proportions of saproxylic beetle populations that utilise clear-
452 cut stumps in a boreal landscape – biodiversity implications for stump harvest. *For. Ecol. Manage.*
453 334, 313-320.

454

455 Joseph, G., Kelsey, R. G., Peck, R.W. and Niwa, C. G., 2001. Response of some scolytids and their
456 predators to ethanol and 4-allylanisole in pine forests of central Oregon. *J. Chem. Ecol.* 27, 699–
457 715.

458

459 Kärhä, K., 2012. Comparison of two stump-lifting heads in final felling Norway spruce stand. *Silva*
460 *Fennica.* 46, 625–640.

461

462 Koistinen, A., Luiro, J-P. and Vanhatalo, K. (toim.), 2016. Metsänhoidon suositukset energiapuun
463 korjuuseen, työopas. Tapion julkaisuja. Available at: [http://www.metsanhoitosuosituksset.fi/wp-](http://www.metsanhoitosuosituksset.fi/wp-content/uploads/2017/05/Metsanhoidon_suosituksset_energiapuun_korjuuseen_Tapio_2016_C.pdf)
464 [content/uploads/2017/05/Metsanhoidon_suosituksset_energiapuun_korjuuseen_Tapio_2016_C.pdf](http://www.metsanhoitosuosituksset.fi/wp-content/uploads/2017/05/Metsanhoidon_suosituksset_energiapuun_korjuuseen_Tapio_2016_C.pdf)
465 (Referred 23th October, 2017)

466

467 Långström, B., 1982. Abundance and seasonal activity of adult *Hylobius*-weevils in reforestation
468 areas during first years following final felling. *Communicationes Instituti Forestalis Fenniae.*

469 106, 1–23.

470

471 Långström, B., and Day, K.R., 2004. Damage, control and management of weevil pests, especially
472 *Hylobius abietis*. Chapter 19, in: Lieutier, F., Day, K. R., Battisti, A. Grégoire, J.-C. & Evans, H. F.
473 (eds.). Bark and wood boring insects in living trees in Europe: a synthesis. Kluwer Academic
474 Publishers, Dordrecht. ISBN 1-4020-2240-9.

475

476 Leather, S.R., Day, K.R., and Salisbury, A.N., 1999. The biology and ecology of the large pine
477 weevil, *Hylobius abietis* (Coleoptera: Curculionidae): a problem of dispersal? Bull. Entomol. Res.
478 89, 3–16.

479

480 Lindelöw, Å., 1992. Seedling mortality caused by *Hylastes cunicularius* Er. (Coleoptera,
481 Scolytidae) in *Picea abies* plantations in Northern Sweden. Scand. J. For Res. 7, 387–392.

482

483 Lindelöw, A., Eidmann, H. H., and Nordenhem, H., 1993. Response on the ground of bark beetle
484 and weevil species colonizing conifer stumps and roots to terpenes and ethanol. Journal of Chem.
485 Ecol. 19: 1393–1403.

486

487 Luoranen, J., Viiri, H., Sianoja, M., Poteri, M. and Lappi, J., 2017. Predicting pine weevil risk:
488 Effects of site, planting spot and seedling level factors on weevil feeding and mortality of Norway
489 spruce seedlings. For. Ecol. Manage. 389, 260–271.

490

491 Löf, M., Paulsson, R., Rydberg, D., and Welander, N.T., 2005. The influence of different overstorey

492 removal on planted spruce and several broadleaved tree species: survival, growth and pine weevil
493 damage during three years. *Ann. For. Sci.* 62, 237–244.

494

495 Mäkipää, R., Linkosalo, T., Komarov, A., and Mäkelä, A., 2015. Mitigation of climate change with
496 biomass harvesting in Norway spruce stands: are harvesting practices carbon neutral? *Can. J. For.*
497 *Res.* 45, 217–225.

498

499 Manlove, J.D., Styles, J., and Leather, S.R., 1997. Feeding of the adults of the large pine weevil,
500 *Hylobius abietis* (Coleoptera: Curculionidae). *Eur. J. Entomol.* 94, 53–156.

501

502 Moore, R., Brixey, J. M., and Milner, A. D., 2004. Effect of time of year on the development of
503 immature stages of the Large Pine Weevil (*Hylobius abietis* L.) in stumps of Sitka spruce (*Picea*
504 *sitchensis* Carr.) and influence of felling date on their growth, density and distribution. *J. Appl. Ent.*
505 128, 167–176.

506

507 Nordlander, G., 1987. A method for trapping *Hylobius abietis* (L.) with a standardized bait and its
508 potential for forecasting seedling damage. *Scand. J. For. Res.* 2, 199–213.

509

510 Nordlander, G., Hellqvist, C., Johansson, K., and Nordenhem, H., 2011. Regeneration of European
511 boreal forests: Effectiveness of measures against seedling mortality caused by the pine weevil
512 *Hylobius abietis*. *For. Ecol. Manage.* 262, 2354–2363.

513

514 Nordlander, G., Bylund, H., Örlander, G., and Wallertz, K., 2003. Pine weevil population density
515 and damage to coniferous seedlings in a regeneration area with and without shelterwood. *Scand. J.*
516 *For Res.* 18, 438–448.

517

518 Nordlander, G., Nordenhem, H., and Bylund, H., 1997. Oviposition patterns of the pine weevil
519 *Hylobius abietis*. Entomol. Exp. Appl. 85, 1–9.

520

521 Nordlander, G., Hellqvist, C., Hjelm, K., 2017. Replanting conifer seedlings after pine weevil
522 emigration in spring decreases feeding damage and seedling mortality. Scand. J. For. Res. 32, 60–
523 67.

524

525 Nordenhem, H., 1989. Age, sexual development, and seasonal occurrence of the pine weevil
526 *Hylobius abietis* (L.). J. Appl. Entomol. 108, 260–270.

527

528 Örlander, G., and Nilsson, U., 1999. Effect of reforestation methods on pine weevil (*Hylobius*
529 *abietis*) damage and seedling survival. Scand. J. For. Res. 14, 341–354.

530

531

532 Örlander, G., Nilsson, U., and Nordlander, G., 1997. Pine weevil abundance on clearcuttings of
533 different ages: a 6-year study using pitfall traps. Scand. J. For. Res. 12, 225–240.

534

535 Ortiz, C.A., Hammar, T., Ahlgren, S., Hansson, P.-A., and Stendahl, J., 2016. Time-dependent
536 global warming impact of tree stump bioenergy in Sweden. For. Ecol. Manage, 371, 5–14.

537

538 Rahman, A., Viiri, H., Pelkonen, P., and Khanam, T., 2015. Have stump piles any effect on the pine
539 weevil (*Hylobius abietis* L.) incidence and seedling damage? Glob Ecol and Conv. 3, 424–432.

540

541 Saarinen, V. M., 2006. The effects of slash and stump removal on productivity and quality of
542 forest regeneration operations – preliminary results. *Biomass Bioenergy* 30, 349–356.

543

544 Shevlin, K. D., Hennessy, R., Dillon, A. B., O'Dea, P., Griffin, C. T., and Williams, C. D., 2017.
545 Stump-harvesting for bioenergy probably has transient impacts on abundance, richness and
546 community structure of beetle assemblages. *Agr.For. Ent.* 19, 388–399. doi:10.1111/afe.12218.

547

548 Skrzecz, I., 2017. Insects associated with reforestation and their management in Poland. Chapter 7,
549 pp. 134–168, in: Shields, V.D.C. (Ed.). *Biological Control of Pest and Vector Insects*. Intech.
550 <http://dx.doi.org/10.5772/6645>.

551

552 von Sydow, F., 1997. Abundance of pine weevil (*Hylobius abietis*) damage to conifer seedlings in
553 relation to silvicultural practice. *Scand. J. For. Res.* 12, 157–167.

554

555 Victorsson, J., and Jonsell, M., 2016. Overlooked subterranean saproxylic beetle diversity in clear-
556 cut stumps and its implications for stump extraction. *For. Ecol. Manage.* 371, 59–66.

557

558 Viiri, H., and Luoranen, J., 2017. Deep planting of Norway spruce seedlings: effects on pine
559 weevil feeding damage and growth. *Can. J. For. Res.* 47, 1468–1473.

560

561 Walmsley, J. D., and Godbold, D.L., 2010. Stump Harvesting for Bioenergy - A Review of the
562 Environmental Impacts. *Forestry* 83, 17–38.

563

564 Wallertz, K., Nordenhem, H., and Nordlander, G., 2014. Damage by the pine weevil *Hylobius*
565 *abietis* to seedlings of two native and five introduced tree species in Sweden. *Silva Fennica* 48(4):
566 article id 1188.

567

568 Work, T.T., Andersson, J., Ranius, T., and Hjältén J., 2016. Defining stump harvesting retention
569 targets required to maintain saproxylic beetle biodiversity. *For. Ecol. Manage.* 371, 90–102.

570

571

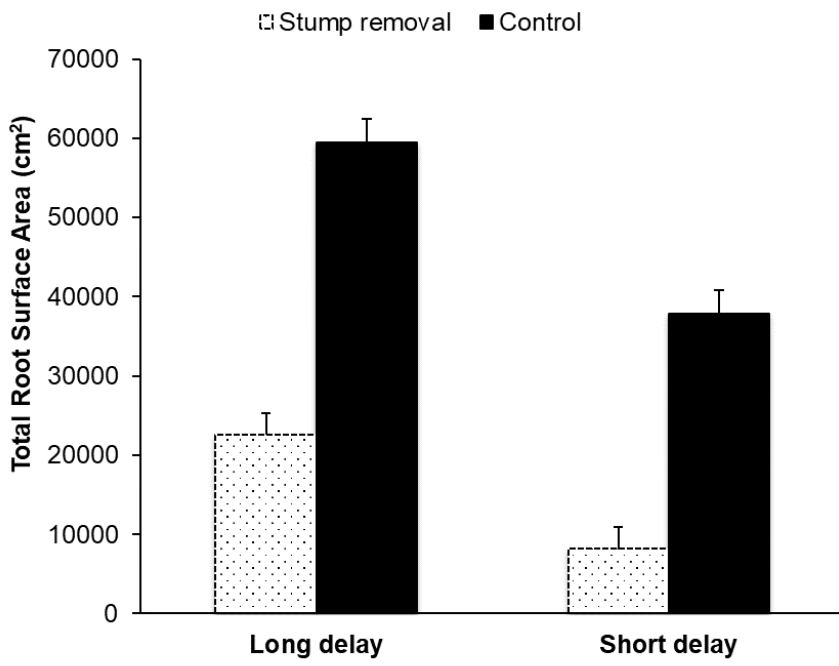
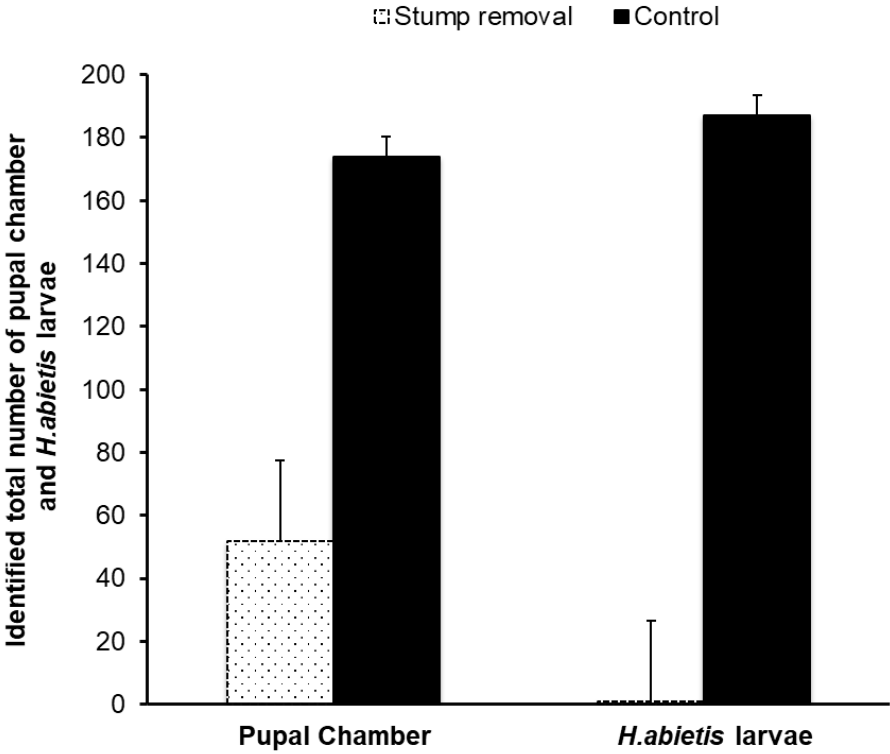
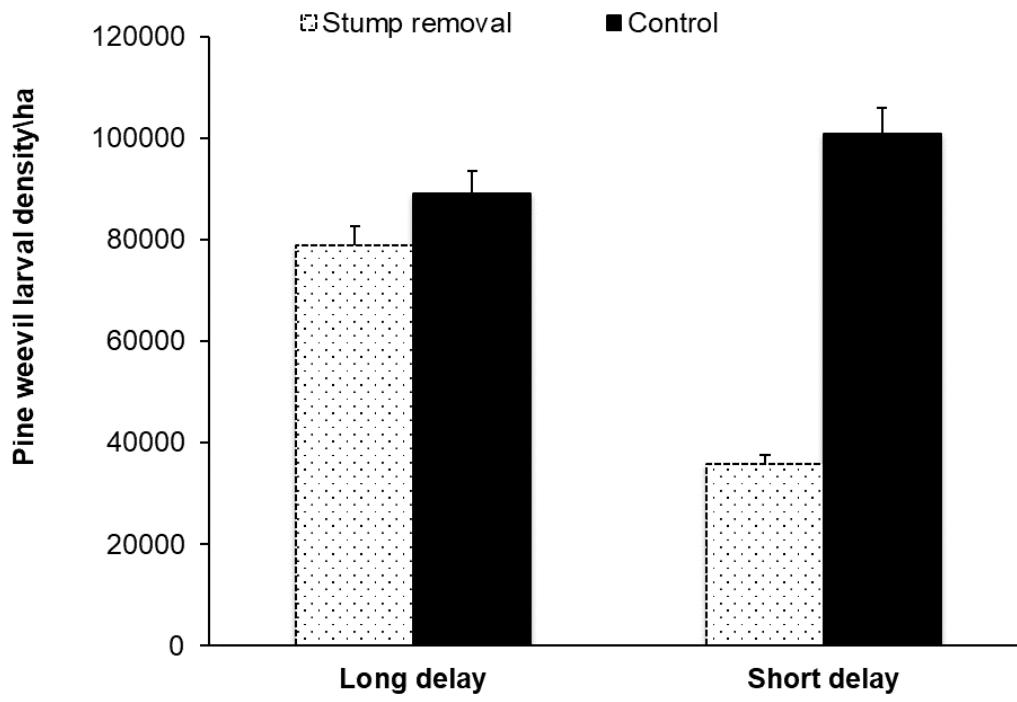


Fig 2





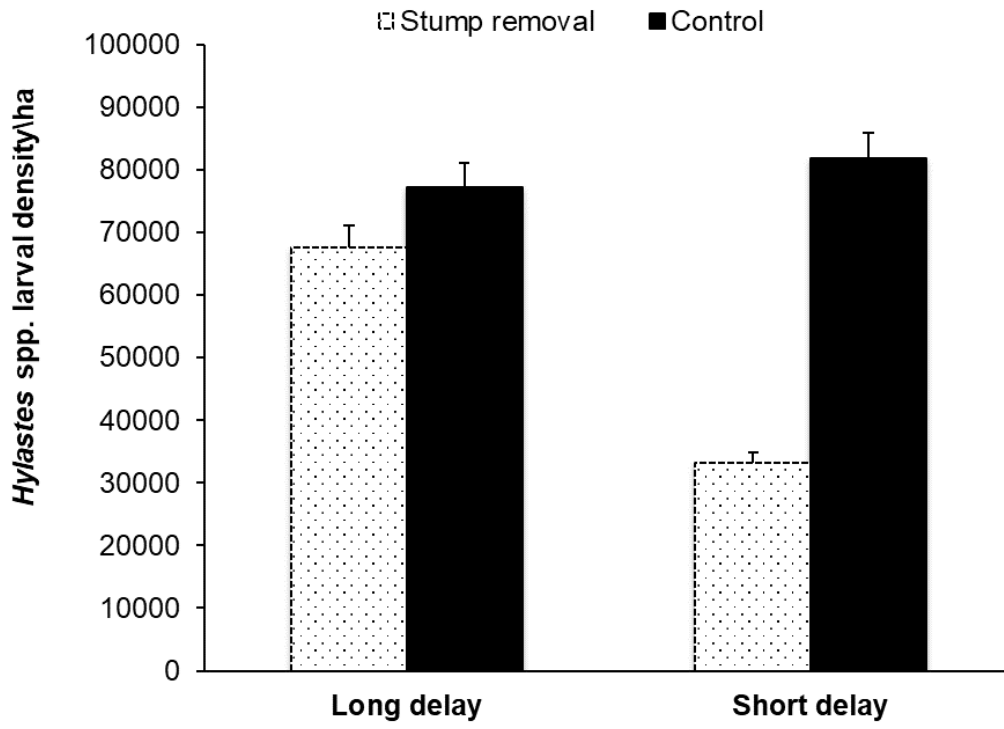
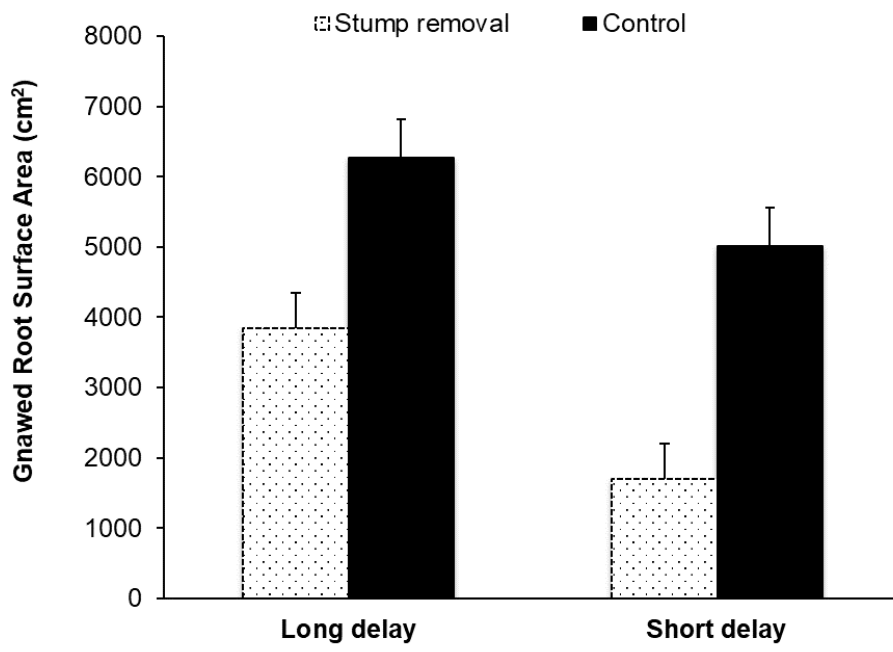


Fig 5



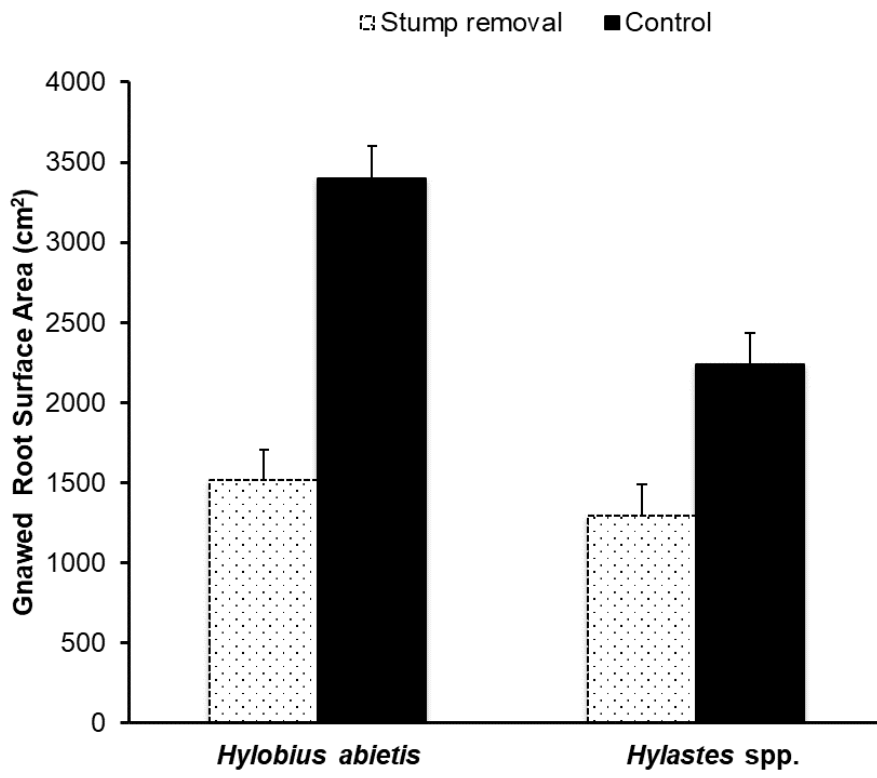


Table 1

Description of study sites and temperature sum, dd °C (threshold > +5 °C).

Site name	Pair code	Distance between pair sites (km)	Logging volume(m ³ /ha)	Coordinates	Area, ha	Site type	dd	Logged	Stump extraction	Root sampling date
Uimaharju	P1	35	223	62°56' 30.24", 30° 19' 28.239"	0.73	Damp (<i>Myrtillus</i> type)	1066	2009	No	9.–20.8.2012
Katajavaara	P1		237	62°51' 33.396", 29° 50' 32.868"	0.89	Damp (<i>Myrtillus</i> type)	1043	2009	2010	8.–16.8.2012
Korpivaara,	P2	55	260	62°50' 22.696", 30° 42' 24.108"	0.91	Damp (<i>Myrtillus</i> type)	1026	2009	No	21.–24.8.2012
Havukkavaara 1	P2		287	62°36' 59.054", 30° 9' 35.772"	2.95	Damp (<i>Myrtillus</i> type)	1017	2009	2010	6.–14.8.2012
Kokonsalmi	P3	85	234	62°26' 2.793", 28° 52' 58.572"	0.70	Damp (<i>Myrtillus</i> type)	1172	2009	No	31.8.–5.9.2012
Havukkavaara 2	P3		244	62°36' 50.369", 30° 9' 39.412"	3.42	Damp (<i>Myrtillus</i> type)	1020	2009	2010	7.–13.8.2012
Rempsu	P4	20	256	62°26' 4.524", 28° 53' 54.41"	1.97	Damp (<i>Myrtillus</i> type)	1167	2009	No	11.–14.9.2012
Juurikka	P4		260	62°32' 3.482", 28° 50' 50.81"	1,12	Damp (<i>Myrtillus</i> type)	1141	2009	2010	31.8.–5.9.2012

Petrumansalo	P5	15	294	62°26' 34.754", 28° 53' 22.978"	2.93	Damp (<i>Myrtillus</i> <i>type</i>)	1154	2009	No	20.–25.9.2012
Juurikkajärvi	P5		298	62°32' 10.792", 28° 51' 0.221"	2.07	Damp (<i>Myrtillus</i> <i>type</i>)	1143	2009	2010	17.–19.9.2012
Polvijärvenniemi	P6	30	304	62°24' 44.94", 28° 19' 22.589"	0.92	Damp (<i>Myrtillus</i> <i>type</i>)	1215	2010	2010	4.–9.9.2013
Jalaslampi	P6		301	62°23' 10.059", 28° 47' 24.971"	1.19	Rich (<i>Oxalis-Myrtillus</i> <i>type</i>)	1179	2010	No	27.9.–2.10.2013
Polvijärvensalmi	P7	30	262	62°24' 45.805", 28° 19' 9.735"	2.23	Damp (<i>Myrtillus</i> <i>type</i>)	1215	2010	2010	11.–13.9.2013
Kermansalo	P7		258	62°23' 39.13", 28° 47' 31.512"	0.95	Rich (<i>Oxalis-Myrtillus</i> <i>type</i>)	1173	2010	No	19.–26.9.2013
Valkeinen	P8	40	267	62°23' 14.925", 28° 48' 46.041"	0.43	Damp (<i>Myrtillus</i> <i>type</i>)	1169	2010	No	9.–14.10.2013
Arhinmäki	P8		265	62°36' 34.767", 28° 39' 37.257"	3.43	Damp (<i>Myrtillus</i> <i>type</i>)	1122	2010	2010	15.–18.10.2013

Table 2

Summary of the different root parameters in control, short delay and long delay stump extraction plots

	Control	Short delay stump extraction	Long delay stump extraction
Number of roots/ha (\pm S.E.)	322 900 \pm 1.11	133 600 \pm 0.88	292 700 \pm 1.60
Diameter of roots cm/ ha, mean(\pm S.E.)	10.4 \pm 0.35	6.7 \pm 0.25	6.8 \pm 0.19
Surface area of roots cm ² / ha, mean(\pm S.E.)	51 546 \pm 2615.9	8 001 \pm 658.1	22 412 \pm 1426.7

Table 3

Results of general linear mixed model analysis of the effect of stump removal on root surface area in soils of forest regeneration sites. Parameters of the model were Treatment (Control, Stump removal), Time lag of stump extraction (Long, short), their interaction term Treatment*Time lag and Logging Volume of total root surface area as continuous covariate of the model

	Df	F-value	P-value
Treatment	11	170.2	0.001
Time lag	11	53.4	0.001
Treatment*time lag	11	9.87	0.009
Logging volume	11	0.58	0.461

Table 4

Results of general linear mixed model analysis of the effect of stump removal on larvae density of *Hylobius abietis* and *Hylastes* spp. Parameters of the model were Treatment (Control, Stump removal), Time lag of stump extraction (Long, short), their interaction term Treatment*Time lag and Logging Volume of total root surface area as continuous covariate of the model

	Pine weevil larvae			<i>Hylastes</i> spp. larvae		
	Df	F-value	P-value	Df	F-value	P-value
Treatment (Control, Stump removal)	11	5,6	0.002	11	9,8	0.009
Time lag (Long, short)	11	17	0.108	11	2,2	0.164
Treatment*time	11	3.06	0.014	11	3,8	0.076
Logging volume	11	0.006	0.940	11	0.001	0.970

Table 5

Results of general linear mixed model analysis of the effect of stump removal on total root surface gnawed area by *Hylobius abietis* and *Hylastes* spp. Parameters of the model were Treatment (Control, Stump removal), Time lag of stump extraction (Long, short), their interaction term Treatment*Time lag and Logging Volume of total root surface area as continuous covariate of the model

Total Root Gnawed area			
	Df	F-value	P-value
Treatment (Control, Stump removal)	6.2	32.48	0.001
Time lag (Long, short)	6.6	3.8	0.096
Treatment*time	6	6.1	0.049
Logging volume	10.4	0.6	0.452

Table 6

Results of general linear mixed model analysis of the effect of stump removal on root surface gnawed area separately by *Hylobius abietis* and *Hylastes* spp. Parameters of the model were Treatment (Control, Stump removal), Time lag of stump extraction (Long, short), their interaction term Treatment*Time lag and Logging Volume of total root surface area as continuous covariate of the model

	Root surface area gnawed by Pine weevil			Root surface area gnawed by <i>Hylastes</i> spp.		
	Df	F-value	P-value	Df	F-value	P-value
Treatment (Control, Stump removal)	6.1	21	0.004	5.4	14.6	0.011
Time lag (Long, short)	6.6	3.1	0.125	5.9	4.3	0.084
Treatment*time	6.1	3.8	0.099	5.5	2.6	0.160
Logging volume	9.9	0.78	0.398	9.9	0.66	0.435